

Fishery Data Series No. 09-17

**Stock Assessment and Restoration of the Afognak
Lake Sockeye Salmon Run, 2007**

by

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March 2009

Alaska Department of Fish and Game

Divisions of Sport Fish and Commercial Fisheries



Symbols and Abbreviations

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Weights and measures (metric)		General		Measures (fisheries)	
centimeter	cm	Alaska Administrative Code	AAC	fork length	FL
deciliter	dL	all commonly accepted abbreviations	e.g., Mr., Mrs., AM, PM, etc.	mid-eye to fork	MEF
gram	g	all commonly accepted professional titles	e.g., Dr., Ph.D., R.N., etc.	mid-eye to tail fork	METF
hectare	ha	at	@	standard length	SL
kilogram	kg	compass directions:		total length	TL
kilometer	km	east	E		
liter	L	north	N	Mathematics, statistics	
meter	m	south	S	<i>all standard mathematical signs, symbols and abbreviations</i>	
milliliter	mL	west	W	alternate hypothesis	H _A
millimeter	mm	copyright	©	base of natural logarithm	<i>e</i>
		corporate suffixes:		catch per unit effort	CPUE
Weights and measures (English)		Company	Co.	coefficient of variation	CV
cubic feet per second	ft ³ /s	Corporation	Corp.	common test statistics	(F, t, χ^2 , etc.)
foot	ft	Incorporated	Inc.	confidence interval	CI
gallon	gal	Limited	Ltd.	correlation coefficient (multiple)	R
inch	in	District of Columbia	D.C.	correlation coefficient (simple)	r
mile	mi	et alii (and others)	et al.	covariance	cov
nautical mile	nmi	et cetera (and so forth)	etc.	degree (angular)	°
ounce	oz	exempli gratia (for example)	e.g.	degrees of freedom	df
pound	lb	Federal Information Code	FIC	expected value	<i>E</i>
quart	qt	id est (that is)	i.e.	greater than	>
yard	yd	latitude or longitude	lat. or long.	greater than or equal to	≥
		monetary symbols (U.S.)	\$, ¢	harvest per unit effort	HPUE
Time and temperature		months (tables and figures): first three letters	Jan, ..., Dec	less than	<
day	d	registered trademark	®	less than or equal to	≤
degrees Celsius	°C	trademark	™	logarithm (natural)	ln
degrees Fahrenheit	°F	United States (adjective)	U.S.	logarithm (base 10)	log
degrees kelvin	K	United States of America (noun)	USA	logarithm (specify base)	log ₂ , etc.
hour	h	U.S.C.	United States Code	minute (angular)	'
hour	h	U.S. state	use two-letter abbreviations (e.g., AK, WA)	not significant	NS
minute	min			null hypothesis	H ₀
second	s			percent	%
Physics and chemistry				probability	P
all atomic symbols				probability of a type I error (rejection of the null hypothesis when true)	α
alternating current	AC			probability of a type II error (acceptance of the null hypothesis when false)	β
ampere	A			second (angular)	"
calorie	cal			standard deviation	SD
direct current	DC			standard error	SE
hertz	Hz			variance	
horsepower	hp			population	Var
hydrogen ion activity (negative log of)	pH			sample	var
parts per million	ppm				
parts per thousand	ppt, ‰				
volts	V				
watts	W				

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ABSTRACT

Beginning in 2001 the Afognak Lake sockeye salmon *Oncorhynchus nerka* runs substantially declined. Concerns expressed by local subsistence users to the Alaska Department of Fish and Game and the US Fish and Wildlife Service Office of Subsistence Management prompted a seasonal investigation of the lake's rearing environment beginning in 2003, which has continued until the present. This report provides the 2007 fishery and limnology results from the Afognak Lake study and fulfills annual reporting requirements to the US Fish and Wildlife Service Office of Subsistence Management, the funding agent for this project (project 07-401).

During 2007, 55,315 sockeye salmon smolt were captured using a Canadian fan trap operated from 10 May to 4 July. Using mark-recapture techniques, we estimated that 275,450 sockeye salmon smolt (95% CI 240,388 – 310,512) emigrated from Afognak Lake. The population was composed of 237,383 age-1. and 38,067 age-2. smolt. Age-1. smolt had a mean weight of 2.6 g, a mean length of 70.4 mm, and a mean condition factor of 0.75. Age-2. smolt had a mean weight of 3.4 g, a mean length of 76.5 mm, and a mean condition factor of 0.74.

Five limnology surveys were conducted in Afognak Lake from May to September, 2007. Seasonal physical conditions and water chemistry values were generally consistent with historical data collected from Afognak Lake; however phosphorus concentrations in 2007 were the lowest recorded levels of phosphorus in Afognak Lake. Seasonal zooplankton density averaged 112,480 animals per m², and cladocerans comprised 64.2% of the zooplankton sampled. The cladoceran *Bosmina* was the most abundant zooplankter, while *Epischura* was the most abundant copepod.

Key words: Afognak Lake, Afognak Island, age, emigration, escapement, Kodiak Island, *Oncorhynchus nerka*, smolt, sockeye salmon, subsistence harvest, trap, zooplankton.

INTRODUCTION

Afognak Lake sockeye salmon *Oncorhynchus nerka* runs substantially declined in 2001 and subsequent escapements from 2002 through 2004 were well below the established sustainable escapement goal (SEG) range of 40,000 to 60,000 sockeye salmon (Baer et al. 2007; Caldenty 2007; Dinnocenzo et al. 2007). As a result of these poor runs, the commercial sockeye salmon fishery in Afognak Bay was closed in 2001 and commercial fishing remained closed until 2005 when a five day opening occurred and 356 fish were harvested (Figure 1). The sockeye salmon commercial fishery remained closed in 2006 and 2007. Sport fishing restrictions were also implemented in 2001, and in-season closures and reduced bag limits have occurred each year through 2004. The sport fishery remained open throughout the 2005 and 2006 seasons without any restrictions but was closed again in 2007. In conjunction with commercial and sport fishing closures, State and Federal managers closed subsistence fishing in early June during the 2002 season, and in-season closures have occurred each year through 2004 in an attempt to achieve the escapement goals for sockeye salmon into Afognak Lake. The subsistence fishery remained open throughout the 2005 and 2006 seasons with minimal harvests while a closure occurred in 2007 through the month of July. Although the subsistence fishing closures restricted harvest of sockeye salmon and caused fishing efforts to shift to other systems, subsistence salmon fishing has been allowed every year in Afognak Bay for pink *O. gorbuscha* and coho *O. kisutch* salmon starting 1 August.

The Afognak Lake sockeye salmon run has historically provided for the largest subsistence salmon fishery on Afognak Island and the second largest in the Kodiak Archipelago (Baer et al. 2007). Local villagers from Port Lions and Ouzinkie as well as Kodiak area residents have traditionally harvested fish in Afognak Bay (Figure 1). The subsistence fishery is prosecuted within the boundaries of the Alaska Maritime National Wildlife Refuge. Subsistence harvests in Afognak Bay from 1990 to 2007 have ranged from 417 (2007) to 12,412 (1997) sockeye salmon (Table 1). The smallest annual sockeye salmon subsistence numbers on record are from the most recent six years (2002-2007).

After Afognak Lake experienced poor runs and fisheries closures in 2002, local subsistence users, represented by the Kodiak-Aleutian Islands Regional Advisory Council, Kodiak Fish and Game Advisory Committee, and Kodiak Tribal Council, contended that a continued closure of the Afognak system would make it more difficult for local residents to harvest sockeye salmon and would shift fishing effort to small sockeye salmon runs in the area and the Buskin River. The Regional Advisory Council, Kodiak Advisory Committee, and Kodiak Tribal Council informed the Alaska Department of Fish and Game (ADF&G) and U.S. Fish and Wildlife Service that the Afognak Lake sockeye salmon run failure constituted an emergency situation for their constituents. In response to this problem, the ADF&G received funding through the Office of Subsistence Management (OSM), Fishery Resources Monitoring Program to determine the feasibility of estimating sockeye salmon smolt production from Afognak Lake. This initial feasibility study, conducted in 2003, showed that sockeye salmon smolt could be effectively trapped in Afognak River and their abundance reliably estimated using mark-recapture techniques (Honnold and Schrof 2004). The ADF&G had little information on juvenile sockeye salmon during their freshwater life history stage, when sockeye salmon mortality rates are usually greatest (Burgner 1991). Thus, smolt abundance studies are important in that they assess survival during the entire freshwater rearing stage, encompassing egg deposition to subsequent smolt emigration.

In addition to smolt production estimates, ADF&G considered it important to collect limnology data to determine the smolt production capacity of Afognak Lake. The ADF&G and Kodiak Regional Aquaculture Association (KRAA) had fertilized (1990-2000) and stocked juveniles (1992, 1994, 1996-1998) into Afognak Lake to enhance the sockeye salmon run (White et al. 1990). As part of the evaluation process, limnological data (phosphorus, nitrogen, chlorophyll a, and zooplankton) were collected three years prior to, during, and three years after fertilization activities. The limnology sampling program was scheduled to end after 2003, unless the ADF&G obtained additional funding to continue collecting limnological data to examine factors that would limit sockeye salmon production in the freshwater rearing environment. Based on the findings from the 2003 feasibility study, the OSM provided funding for a three-year study (2004-2006) that enabled the continuation of smolt assessment work, examination of rearing and spawning capacity, and estimation of the sockeye salmon production potential of Afognak Lake. A final report in 2007 consolidated historical fishery and limnological data, provided results of a sockeye salmon escapement goal review and production analysis conducted in 2004-2005, and documented the final results of the project. The three year study indicated that rearing conditions within Afognak Lake appeared to be stable or improving and zooplankton abundance did not suggest overgrazing. Favorable rearing conditions were also reflected in the relatively high condition factor of the smolt (>0.75) that enabled most juveniles (86%) to emigrate at age-1.

Continued analysis of Afognak Lake and annual smolt emigration studies were determined to be of high importance to evaluate if there were changes in the nutrient-food web dynamics (e.g., if the structure of consumer communities have modified nutrient transfer along the food web) and how these changes may have affected the growth and production of the juvenile sockeye salmon emigrating from Afognak Lake. Recognizing the importance of continued analysis on Afognak Lake sockeye salmon production, the OSM extended funding to ADF&G for an additional three-year study (2007-2009). This annual report summarizes the 2007 fishery and limnological results associated with the Afognak Lake system.

OBJECTIVES OF THE PROJECT

1. Estimate the number, age, and average size at age of sockeye salmon smolt emigrating from Afognak Lake for 2007 through 2009.
 - Estimate the number (achieving 25% relative error) with a 95% confidence.
 - Estimate the age proportion within $d=0.03$ of the true proportion with 95% confidence.
 - Estimate the average length within 0.5 mm of the true average length and the average weight within 0.25 g of the true average weight with a 95% confidence.
2. Evaluate the water chemistry, nutrient status, and plankton production of Afognak Lake from 2007-2009.
3. Assess the rearing conditions for juvenile sockeye salmon in Afognak Lake based upon completion of objectives 1 and 2.

BACKGROUND

A counting weir was first established just below the lake outlet on the upper reaches of the Afognak River in 1921 and was operated intermittently through 1977. Since 1978 to the present, escapement data has been collected annually. In 1986, the weir was relocated to its current location, just before terminating into the bay, and ADF&G has maintained annual weir counts in conjunction with sockeye salmon age, length and sex (ALS) sampling. Catch data have been documented through the ADF&G commercial landing fish ticket system, statewide sport fish surveys, and return of subsistence fishing permits since the late 1970s (Dinnocenzo et al. 2007).

Prior to 2005 the Afognak Lake sockeye salmon escapement goal range was 40,000 to 60,000 fish (Nelson and Lloyd 2001). Escapements in 1987 and 1988 did not reach the lower end of the range, and little commercial fishing effort was directed at this stock through the mid to late 1980s (White et al. 1990). In the mid 1980s, Kodiak Island residents surveyed by the Kodiak Regional Planning Team (KRPT) indicated that sockeye salmon were the preferred species for commercial and subsistence fishers in the area (KRPT 1987). These results, coupled with the declining sockeye salmon production from Afognak Lake, resulted in the system being listed by the KRPT and KRAA as the highest priority sockeye salmon enhancement project on Afognak Island. In 1987, the ADF&G, in cooperation with KRAA, initiated pre-fertilization fisheries and limnological investigations at Afognak Lake (Honnold and Schrof 2001; Schrof et al. 2000; White et al. 1990). Results of these investigations indicated that sockeye salmon production was limited by rearing capacity (White et al. 1990). Nutrient enrichment was recommended and then implemented in 1990 to increase primary and secondary production, which was intended to increase sockeye salmon rearing capacity in the lake.

Adult sockeye salmon from Afognak Lake were screened for disease in 1987 and 1988 as part of an evaluation of the stock as a candidate for an early-run brood source for future KRAA enhancement projects (Schruf et al. 2000; White et al. 1990). The Afognak Lake sockeye salmon stock was selected as a brood stock for barren lake stocking projects on Afognak Island, with the first fish stocked in Little Waterfall, Hidden, and Crescent Lakes in 1992 (McCullough et al. 2000; McCullough et al. 2001; McCullough and Clevenger 2002). Hatchery survivals were higher than anticipated in 1992 and resulted in more fry being available than had been planned. Rather than increasing stocking levels into the barren lakes, which had not been stocked previously, the ADF&G allowed KRAA to stock the excess fry back into Afognak Lake. Although the escapement in 1992 (and from 1989 to 1991) exceeded the sustainable escapement

goal, stocking a fairly small number of juveniles (less than 500,000) was considered acceptable as long as the lake fertilization program continued and zooplankton (primary forage for juvenile sockeye) levels remained stable. Afognak Lake stocking was repeated in 1994, and from 1996 to 1998. To alleviate concerns of increasing the grazing pressure exerted by stocked fry on the zooplankton population, lake fertilization was continued. In 1999, the ADF&G wanted the KRAA to follow the established egg take goals in order to avoid stocking excess fry into Afognak Lake (Honnold et al. 1999). The number of sockeye salmon eggs that could be taken from Afognak Lake by KRAA was reduced, and fertilization of Afognak Lake was also discontinued after 2000 (Honnold and Schrof 2001).

Beginning in 2000, the Alaska Board of Fisheries adopted two policies into regulation to ensure that the state's salmon stocks would be conserved, managed and developed using the sustained yield principle. In 2000 the Policy for the Management of Sustainable Salmon Fisheries (5 AAC 39.222) was adopted and in 2001 the Policy for Statewide Salmon Escapement Goals (5 AAC 39.223) was put into regulation.

Two important terms defined in the Policy for the Management of Sustainable Salmon Fisheries are:

“Biological escapement goal (BEG): the escapement that provides the greatest potential for maximum sustained yield (MSY)” and,

“Sustainable escapement goal (SEG): a level of escapement, indicated by an index or an escapement estimate, that is known to provide for sustained yield over a 5 to 10 year period, used in situations where a BEG cannot be estimated due to the absence of a stock-specific catch estimate.”

In 2004, using these new policies, a team of ADF&G biologists re-evaluated the existing Afognak Lake sockeye salmon escapement goal. The team recommended changing the escapement goal from a SEG of 40,000 to 60,000 sockeye salmon (Nelson and Lloyd 2001) to a BEG of 20,000 to 50,000 sockeye salmon (Nelson et al. 2005). The recommendation was based on analysis of spawner recruit and limnology data, excluding data from years in which the lake was fertilized. In January 2005, the Directors of Commercial Fisheries and Sport Fish Divisions approved these recommendations. In 2007, the escapement goal was reevaluated with three additional years of data and was recommended to remain a BEG of 20,000 to 50,000 fish (Honnold et al. 2007). Escapements during the last seven years have been just below (2002 and 2004) to just above (2001, 2003, 2005-2007) the lower end of the new BEG range (Table 1). However, the Policy for Sustainable Salmon Management instructs the ADF&G “to maintain evenly distributed salmon escapements within the bounds of the BEG.”

Juvenile production studies have been conducted in conjunction with limnological investigations at a number of sockeye salmon systems in the Kodiak archipelago (Barrett et al. 1993a, 1993b; Coggins 1997; Coggins and Sagalkin 1999; Edmundson et al. 1994a, 1994b; Honnold 1997; Honnold and Edmundson 1993; Kyle et al. 1988, 1990; Kyle and Honnold 1991; Sagalkin 1999; Sagalkin and Honnold 2003; Schrof et al. 2000; Swanton et al. 1996; White et al. 1990). Some of these studies estimated smolt abundance and size by age through trapping and mark-recapture techniques. Several studies also counted the entire smolt emigration by use of a weir and trap. Rearing juveniles in lakes were enumerated using hydroacoustics and trawl surveys. Smolt abundance and size studies provide estimates of overall freshwater survival, covering the time between egg deposition in the gravel and smolt emigration to the ocean.

Prior to 2003, ADF&G efforts to collect juvenile sockeye salmon data from Afognak Lake had met with limited success (Schrof et al. 2000). Estimates of lake rearing juveniles using hydroacoustics proved inaccurate due to the presence of large numbers of sticklebacks. Due to difficulties associated with species separation, the hydroacoustic surveys were discontinued after 1995. Smolt abundance data were collected in 1990 and 1991, but reliable smolt estimates were not obtained due to low trap efficiencies identified during mark-recapture trials, which were probably caused by poor trap design. In 1992, funding for the mark-recapture project was discontinued and only smolt age, weight, and length data collection was continued. Further funding reductions resulted in smolt age, weight, and length data collection being limited to one annual sample after 1995. It was not until 2003 that smolt abundance and age composition estimates again became available with implementation of a feasibility study under the Fishery Resource Monitoring Program (Honnold and Schrof 2004).

In addition to smolt abundance and size data, additional information on the rearing conditions within Afognak Lake is needed to determine what other factors may be affecting sockeye salmon production. Such work is essential because sockeye salmon mortality rates are highest during the freshwater life history stage (Burgner 1991). A lake's physical parameters (solar illumination, temperature, and dissolved oxygen) greatly affect nutrient cycling (Schlesinger 1991). Lake nutrients, specifically phosphorous and nitrogen, are prerequisites for photosynthesis; and their concentrations can be used to assess the potential for primary production within a system (Spalinger and Bouwens 2003). Chlorophyll *a* levels are indicators of the standing crop of primary producers that provide food for zooplankton, which are prey for sockeye salmon. Estimating zooplankton population attributes is crucial to understanding the progression of a lacustrine food chain. Zooplankton abundance, individual size, and species composition can be regulated from the bottom-up by phytoplankton availability (Stockner and MacIsaac 1996), or from the top-down by predation pressures such as grazing by juvenile sockeye salmon (Kyle 1992). Fortunately, historical limnology data for Afognak Lake exists and includes periods when the sockeye salmon stock was abundant. Comparisons with data collected under current stock conditions could further elucidate potential causes for the decline of sockeye salmon production within Afognak Lake.

Sockeye salmon freshwater production is also limited by the amount and quality of available spawning habitat (Honnold and Edmundson 1993; Willette et al. 1995). In 2005, spawning habitat surveys were conducted on the streams and tributaries of Afognak Lake (Baer et al. 2007). Estimating available lake shore spawning habitat was attempted, but has proven to be difficult to assess.

DESCRIPTION OF STUDY AREA

The Afognak Lake system is located on the southeast side of Afognak Island, approximately 50 km northwest of the city of Kodiak (Figure 1). The Afognak Native Corporation owns the land surrounding the Afognak Lake system down to tidewater. Afognak Lake (58° 07' N, 152° 55' W) lies 21.0 m above sea level, is 8.8 km long, has a maximum width of 0.8 km, and has a surface area of 5.3 km² (Schrof et al. 2000; White et al. 1990). The lake has a mean depth of 8.6 m, a maximum depth of 23.0 m, and a water residence time of 0.4 years (Figure 2). Runoff from Afognak Lake flows in an easterly direction into the 3.2 km long Afognak River, which in turn flows into Afognak Bay, which is part of the Alaska Maritime National Wildlife Refuge and where most subsistence fishing occurs.

In addition to sockeye salmon, resident fish in the Afognak Lake drainage include pink salmon, coho salmon, rainbow trout (anadromous and non-anadromous) *O. mykiss*, Dolly Varden *Salvelinus malma*, three spine stickleback *Gasterosteus aculeatus*, and coastrange sculpin *Cottus aleuticus* (White et al. 1990). Chinook *O. tshawytscha* and chum *O. keta* salmon have been observed in the Afognak River on occasion, but have not established viable spawning populations (White et. al 1990).

METHODS

SMOLT ASSESSMENT

Trap Deployment and Assembly

An inclined-plane Canadian fan trap (Ginetz 1977; Todd 1994) was installed on 10 May 2007 approximately 32 m upstream from the confluence of the Afognak River and Afognak Bay. The trap was positioned towards the middle of the river, where water velocity was great enough to make it difficult for smolt to avoid capture (Figure 3). A live box (1.2 m x 1.2 m x 0.5 m) was attached to the cod end of the trap, and the entire trapping device was suspended from a cable attached to a come-along. The trap was secured to an aluminum pipe frame, which allowed the vertical trap position to be adjusted in response to water level fluctuations. Perforated (3.2 mm) aluminum sheeting (1.2 m x 2.4 m) supported by a Rackmaster®¹ pipe frame was placed at the entrance of the trap in a “V” configuration to divert smolt into the live box. Trapping ceased and the trap was removed from the river on 4 July, after smolt abundance declined and the number captured was less than 100 smolt per day for three consecutive days. Detailed methods for trap installation, operation, and maintenance are described in Baer (2007).

Smolt Enumeration

Smolt were captured in the trapping system and held in the attached live box until they were counted. During the evening (2200 to 0800 hours), the live box was checked every one to two hours, depending on smolt abundance. During the day (0801 to 2159 hours), the live box was checked every three to four hours. All smolt were removed from the live box with a dip net, counted, and either released downstream of the trap or transferred to an in-stream holding box for sampling and marking. Species identification was made by visual examination of external characteristics (Pollard et al. 1997). All data, including mortality counts, were entered on a reporting form each time the trap was checked.

Trap Efficiency and Population Estimates

Mark-recapture experiments were performed to measure smolt trap efficiency (e). Sockeye salmon smolt were collected, marked with Bismark Brown Y dye, and released about once per week as well as when changes were made to the trapping system. Based on smolt studies at Akalura Lake (Coggins and Sagalkin 1999; Sagalkin and Honnold 2003), we attempted to achieve trap efficiencies between 15 to 20%. To estimate the desired trap efficiency and be within the relative error (r) of 25% in estimating total abundance, we needed to mark and release 300 to 500 smolt for each experiment (Carlson et al. 1998; Robson and Regier 1964;). Once collected, smolt were placed in an aerated 33-gallon trashcan filled with water and transported, in a trailer pulled by an all-terrain vehicle to the release site approximately 1,240 m upstream. At

¹ Product names used in this report are included for scientific completeness, but do not constitute a product endorsement.

the release site, smolt were exposed to a continuously oxygenated solution of Bismark Brown Y dye (1.9 g of dye to 15 gallons of water) for 30 minutes. The smolt were then transferred to a holding box at the release site. Between 2100 and 2300 hours, most of the dyed smolt (~400) were randomly selected from the holding box, counted, and released across the width of the stream. The remaining dyed smolt (~100) were counted and left in the holding box for five days to estimate delayed mortality resulting from the capture and marking process. Dyed smolt from both groups that displayed unusual behavior (labored breathing, flared gills, side swimming, etc.) were removed from the experiment and released downstream of the trap. The proportion of smolt that died during the five day holding period was used to estimate the actual number marked smolt available for recapture in the experiment (M_h). All dyed smolt recaptured at the trap site were counted and assigned to a capture period and corresponding trap efficiency trial (hereafter referred to as a stratum).

Trap efficiency for each stratum (h) was calculated by dividing the total number of dyed smolt recaptured by the number of dyed smolt released within the stratum:

$$e_h = \frac{m_h}{M_h} \quad (1)$$

where,

- e_h = trap efficiency or smolt capture probability in stratum h ;
- M_h = number of marked smolt released in stratum h and adjusted for estimated delayed mortality;
- m_h = number of marked smolt recaptured in stratum h .

A modification of the stratified Peterson estimator (Carlson et al. 1998) was used to estimate the number of smolt emigrating within each stratum:

$$\hat{U}_h = \frac{u_h(M_h + 1)}{m_h + 1} \quad (2)$$

where,

- U_h = total number of smolt in stratum h , excluding marked releases and minus observed mortality;
- u_h = number of unmarked smolt recaptured in stratum h ;

Variance of the smolt abundance estimate was calculated as:

$$v(\hat{U}_h) = \frac{(M_h + 1)(u_h + m_h + 1)(M_h - m_h)u_h}{(m_h + 1)^2(m_h + 2)} \quad (3)$$

Within each stratum h , the total population size by age class j is estimated as,

$$\hat{U}_{jh} = \hat{U}_h \hat{\theta}_{jh}, \quad (4)$$

where $\hat{\theta}_{jh}$ is the proportion of age class j during each stratum h . Variance of $\hat{\theta}_{jh}$ was calculated using the standard variance of a population proportion (Thompson 1987). The variance of \hat{U}_{jh} was calculated as,

$$v(\hat{U}_{jh}) = \hat{U}_h^2 v(\hat{\theta}_{jh}) + v(\hat{U}_h) \hat{\theta}_{jh}^2 \quad (5)$$

The total number of emigrating smolt by age class was calculated by summing the individual strata. Variance of the total emigration estimates was calculated by summing the individual variances.

Age, Weight, and Length Sampling

Approximately 200 sockeye salmon smolt were sampled each statistical week to obtain age, weight, and length data. To reach the weekly total, daily samples of 40 sockeye salmon smolt were collected for five days within each statistical week. Smolt were collected throughout the night and held in the in-stream live box. The number of smolt collected each hour was proportional to emigration abundance. Forty smolt were randomly collected from those retained in the live box and sampled to obtain daily age, weight, and length data. After sampling, all smolt were released downstream from the trap.

Tricaine methanesulfonate was used to anesthetize smolt prior to sampling. Fork lengths were measured to the nearest 1 mm, and weights were recorded to the nearest 0.1 g. Scales were removed from the preferred area (INPFC 1963) and mounted on a microscope slide for age determination. After sampling, smolt were held in aerated buckets of water until they recovered from the anesthetic, and subsequently released downstream from the trap. Age was estimated from scales viewed with a microfiche reader at 60X magnification, and recorded in European notation (Koo 1962).

Condition factor (Bagenal and Tesch 1978), a quantitative measure of “fatness,” was determined for each sampled smolt as:

$$K = \frac{W}{L^3} 10^5 \quad (6)$$

where,

- K = smolt condition factor;
- W = weight in g;
- L = snout to fork length in mm.

Life History-Based Population Estimates

We also estimated the number of smolt expected to emigrate in 2007 using Afognak Lake escapement data along with actual fecundity estimates and egg to smolt survival rates reported from other clear water lake systems (Bradford 1995; Drucker 1970; Koenings and Kyle 1997). This alternative method of estimating Afognak Lake smolt emigration incorporated sockeye salmon escapement data, female fecundity data, egg-to-smolt survival estimates, and age composition data to generate a theoretical smolt production estimate by year. Using parent spawning escapements in 2004 and 2005, we assumed a 1:1 sex ratio, an average egg deposition of 2,112 per female (average number of eggs per female as determined from 2003-2006 fecundity estimates by Pillar Creek Hatchery), 7% egg-to-fry survival (Bradford 1995; Drucker 1970; Koenings and Kyle 1997), and 21% fry-to-smolt survival (Koenings and Kyle 1997).

LIMNOLOGICAL ASSESSMENT

Lake Sampling Protocol

Five limnological surveys of Afognak Lake were conducted at approximately 4-5 week intervals from May to September, 2007. Collected data and water samples were returned to the ADF&G Near Island Laboratory and analyzed as described in Thomsen (2008). Two stations, marked with anchored mooring buoys and located with Global Positioning System (GPS) equipment, were sampled from a float-equipped aircraft during each survey (Figure 2). Zooplankton samples were collected at both stations, but water samples were only collected at Station 1. During each survey, water samples for general chemistry and nutrient analysis were collected at a depth of 1 m below the water surface using a 4-L Van Dorn sampler. Each water sample was emptied into a pre-cleaned polyethylene carboy, which was kept cool and dark in the float of the plane until processed at the ADF&G laboratory in Kodiak. Vertical zooplankton hauls were made at each station using a 0.2 m diameter conical net with 153 μm mesh. The net was pulled manually at a constant speed ($\sim 0.5 \text{ m sec}^{-1}$) from approximately 2 m off the lake bottom to the surface. The contents from each tow were emptied into a 125-ml polyethylene bottle and preserved in 10% neutralized formalin.

Temperature, Dissolved Oxygen, Light, Water Clarity and Euphotic Volume

Water temperature ($^{\circ}\text{C}$) and dissolved oxygen (mg L^{-1}) levels were measured with a YSI® meter. Surface readings were calibrated against a hand-held mercury thermometer. Readings were recorded at half-meter intervals to a depth of 5 m, and then at one-meter depth intervals to the lake bottom. Results were categorized into spring (May-June), summer (July-August), and fall (September-October) sampling periods.

Measurements of photosynthetically active wavelengths (PAR) were taken with a Protomatic® submersible photometer sensitive to the visible spectrum range (400-700 nanometers). Readings were taken above the water surface, at the water surface, and at half-meter intervals below the water surface until reaching a depth of 5 m, and then at one-meter intervals until either the lake bottom or a depth equivalent to 1% of the subsurface reading was reached. The mean euphotic zone depth was determined (Koenings et al. 1987) for each lake and used in a model to estimate sockeye salmon fry production (Koenings and Kyle 1997). The vertical extinction coefficient for downward light (K_d , m^{-1}) was obtained from the relation:

$$I_z = I_0 e^{-K_d z} \text{ or } \ln I_z = \ln I_0 - K_d z \quad (7)$$

where,

- I_0 = light penetration just below the surface (Wetzel and Likens 1991);
- I_z = light penetration at z meters (Wetzel and Likens 1991); and
- K_d = the linear regression coefficient of $\ln I_z$ against depth (z).

Assuming K_d is constant with depth, mean euphotic zone depth, the depth at which 1% of the subsurface light remains, is given by $4.6/K_d$ (Kirk 1994).

One-meter temperature and dissolved oxygen profiles were compared to assess the physical conditions in the euphotic zones of the lake. Secchi disc readings were collected from each station to measure water transparency. The depths at which the disc disappeared when lowered into the water column and reappeared when raised in the water column were recorded and averaged.

Lake primary production potential for rearing juvenile sockeye salmon was assessed through a euphotic volume calculation (Koenings and Burkett 1987; Nelson et al. 2005). To calculate euphotic volume, the average mean euphotic zone depth was multiplied by the surface area (5.3 km²).

General Water Chemistry, Phytoplankton and Nutrients

Unfiltered water was analyzed for total phosphorus (TP), total Kjeldahl nitrogen (TKN), pH, and Alkalinity. Sample water was filtered through a rinsed 4.25 cm diameter Whatman GF/F cellulose fiber filter and stored frozen in phosphate free soap-washed polyethylene bottles. Filtered water was also analyzed for total filterable phosphorus (TFP), filterable reactive phosphorus (FRP), nitrate + nitrite (NO₃⁻ + NO₂⁻), ammonia (NH₄⁺) and reactive silicon.

TP, TFP and FRP were analyzed using a Spectronic Genesys 5 Spectrophotometer (SG5) using the potassium persulfate-sulfuric acid digestion method described in Thomsen (2008). Unfiltered frozen water was sent to South Dakota University for the TKN analysis. The pH of water samples was measured with a Corning 430 meter, while alkalinity (mg L⁻¹ as CaCO₃) was determined from 100 ml of unfiltered water titrated with 0.02 N H₂SO₄ to a pH of 4.5 and measured with a pH meter (Mettler Toledo Seven Easy).

Samples for NO₃⁻ + NO₂⁻ were analyzed using the cadmium reduction method described in Thomsen (2008). NH₄⁺ was analyzed with a SG5 using the phenol-sodium hypochlorite method described in Thomsen (2008). Total nitrogen, the sum of TKN and NO₃⁻ + NO₂⁻, and the ratio of total nitrogen to TP was calculated for each sample.

For chlorophyll *a* (chl *a*) analysis, 1.0 L of water from each sample was filtered through a Whatman GF/F filter under 15 psi vacuum pressure. Approximately 5 ml of magnesium chloride (MgCO₃) were added to the final 50 ml of water near the end of the filtration process to act as a preservative. Filters were stored frozen on individual Plexiglas slides until analyzed. Filters were then ground in 90% buffered acetone using a mortar and pestle, and the resulting slurry was refrigerated in separate 15-ml glass centrifuge tubes for 4 hours to ensure maximum pigment extraction. Pigment extracts were centrifuged, decanted, and diluted to 15 ml with 90% acetone. The extracts were analyzed with a SG5 using methods described in Thomsen (2008).

Reactive Silicon was determined with a SG5 spectrophotometer using the ammonium molybdate – sodium sulfite method described in Thomsen (2008). Total filterable phosphorus was determined using the same methods as those for TP utilizing filtered water. Filterable reactive phosphorus was determined using the potassium persulfate- sulfuric acid method described in Thomsen (2008).

Zooplankton

Cladocerans and copepods were identified to the genus using taxonomic keys in Edmondson (1959). Zooplankton lengths were measured in triplicate 1 ml subsamples taken with a Hansen-Stempel pipette and placed in a Sedgewick-Rafter counting chamber. Lengths from a minimum of 15 animals of each species or group (typically animals are grouped at the genus level) were measured to the nearest 0.01 mm, and averaged. Biomass was estimated from species-specific linear regression equations of length and dry weight derived by Koenings et al. (1987). Zooplankton density and biomass data from the two stations were averaged for each survey.

RESULTS

SMOLT ASSESSMENT

Enumeration and Sampling

Smolt trapping was conducted for a total of 56 days, from 10 May to 4 July, 2007. During this period, 55,315 sockeye salmon smolt were captured (Table 2). The greatest daily sockeye salmon smolt catch was obtained on 6 June when 4,294 smolt were captured (Table 2; Figure 4). During the trapping period a total of 1,090 smolt were collected for biological sampling purposes, of which 1,089 were usable for age, weight, and length data (Table 2).

Trap Efficiency and Population Estimates

Five mark-recapture experiments were conducted during the sockeye salmon smolt emigration period in 2007 (Table 2). Trap efficiencies ranged from 11.2% during the fifth experiment (28 June - 4 July) to 61.5% during the second experiment (6 - 12 June). Mean trap efficiency for the entire emigration was 19.9%. The total number of sockeye salmon smolt emigrating from the Afognak Lake system in 2007 was estimated to be 275,450 (95% CI 240,388 – 310,512; Table 3).

Age, Weight, and Length Sampling

Summing the emigration estimates by age over the strata resulted in an estimated emigration estimate of 237,383 age-1. (86.2%) and 38,067 age-2. (13.8%) smolt (Table 4; Figure 5). Age-1. smolt comprised 69.1% of the sample from the first stratum (10 May - 5 June), the second stratum (6 - 12 June) was composed of 94.0% age-1. smolt, the third stratum (13-20 June) contained 99.6% age 1. fish, and the fourth (21-27 June) and fifth strata (28 June - 4 July) both contained 100.0% Age 1. fish.

The sampled age-1. smolt had a mean weight of 2.6 g, a mean length of 70.4 mm, and a mean condition factor of 0.75 (Table 5). The sampled age-2. smolt had a mean weight of 3.4 g, a mean length of 76.5 mm, and a mean condition factor of 0.74 (Table 5).

Comparisons of Mark-Recapture and Life History-Based Population Estimate Methods

Using the life history-based population estimate method we projected that the 2004 escapement of 15,181 adults (brood year 2004) would produce 32,521 age-2. smolt and the 2005 escapement of 21,577 adults (brood year 2005) would produce 288,722 age-1. smolt (Table 6). A total of 321,243 smolt was projected to emigrate from Afognak Lake in the spring of 2007 (Figure 6).

LIMNOLOGICAL ASSESSMENT

Temperature, Dissolved Oxygen, Light, Water Clarity and Euphotic Volume

In 2007, water temperatures ranged from 6.7° C near the lake bottom in the spring (May-June) to 15.4° C at the surface in summer (July-August). Temperature stratification was most noticeable in the summer when the average surface temperature was 15.4° C and the average bottom temperature was 9.5° C. Surface and bottom temperatures in the spring and fall (September-October) were similar, indicating that mixing occurred throughout the entire water column during the spring and fall sampling periods. Dissolved oxygen concentrations ranged from 6.6 mg L⁻¹ at the bottom in the summer to 11.4 mg L⁻¹ at the surface in the spring.

The mean vertical extinction coefficient ($K_d \text{ m}^{-1}$) or rate of light attenuation was -0.50 m^{-1} in 2007. The mean euphotic zone depth was 9.36 m, while the Secchi disk reading was 4.15 meters. The euphotic volume for Afognak Lake in 2007 was $49.61 \text{ } 10^6 \text{ m}^3$.

General Water Chemistry, Phytoplankton and Nutrients

The pH averaged 6.77 units with little seasonal variation (Table 7). Alkalinity levels (measured as CaCO_3) ranged from 10.0 mg L^{-1} to 12.5 mg L^{-1} and averaged 10.9 mg L^{-1} for the five samples collected. Seasonal chl *a* (phytoplankton) concentrations ranged from $0.96 \text{ } \mu\text{g L}^{-1}$ to $2.24 \text{ } \mu\text{g L}^{-1}$ and averaged $1.47 \text{ } \mu\text{g L}^{-1}$ (Table 7).

Seasonal mean TP concentrations ranged from 3.2 to $3.7 \text{ } \mu\text{g L}^{-1}$ and averaged $3.6 \text{ } \mu\text{g L}^{-1}$ (Table 8). Seasonal inorganic phosphorous concentrations of TFP ranged from $0.6 \text{ } \mu\text{g L}^{-1}$ to $1.6 \text{ } \mu\text{g L}^{-1}$ and averaged $1.1 \text{ } \mu\text{g L}^{-1}$ (Table 8). The FRP concentrations ranged from 0.4 to $1.9 \text{ } \mu\text{g L}^{-1}$ and averaged $1.1 \text{ } \mu\text{g L}^{-1}$.

Nitrogen levels were measured in three forms: TKN, $\text{NO}_3^- + \text{NO}_2^-$, and NH_4^+ . The seasonal mean TKN was $115.0 \text{ } \mu\text{g L}^{-1}$, and the greatest seasonal difference was between the September ($86.0 \text{ } \mu\text{g L}^{-1}$) and May ($178.0 \text{ } \mu\text{g L}^{-1}$) samples (Table 8). Seasonal NH_4^+ levels averaged $5.6 \text{ } \mu\text{g L}^{-1}$ and ranged from 4.6 to $6.5 \text{ } \mu\text{g L}^{-1}$. Seasonal $\text{NO}_2 + \text{NO}_3$ levels averaged $55.5 \text{ } \mu\text{g L}^{-1}$ and had a wide range of variability throughout the season, from 21.3 to $124.5 \text{ } \mu\text{g L}^{-1}$ (Table 8). Total nitrogen concentrations ranged from 118.3 to $302.5 \text{ } \mu\text{g L}^{-1}$ and averaged $170.5 \text{ } \mu\text{g L}^{-1}$. The seasonal total nitrogen to total phosphorus ratio, by weight, averaged 107.7:1 (Table 8).

Zooplankton

Zooplankton weighted mean density was 112,480 animals per m^2 at Afognak Lake (Table 9). All zooplankton identified were crustaceans commonly referred to as either cladocerans (*Order* Anomopoda and Ctenopoda) or copepods (*Order* Calanoida, Cyclopoida, and Harpacticoida). Cladocerans were the dominant zooplankter in samples (64.2% of mean), with the genus *Bosmina* being most abundant (58.4% of mean). The other cladoceran genera included, *Daphnia* (1.9% of mean), *Holopedium* (1.7% of mean), and a group we called “other cladocerans,” which consisted of various unidentified immature cladocera (2.3% of mean). Of the copepods (35.8% of mean), the most abundant group consisted of what we called “other copepods” (15.5% of the mean), which was made up mostly of the genus *Harpacticus* and various unidentified nauplii (larvae), followed in abundance by the genus *Epischura* (12.1% of the mean). The copepod genus *Cyclops*, considered an important member of the zooplankton community in sockeye salmon lakes, were not very abundant (6.4% of mean). The genus *Diaptomus* made up the smallest portion of the copepods at 1.8% of the mean.

Zooplankton mean biomass was 127.9 mg per m^2 (Table 9). Despite only making up 35.8% of the mean density the copepods comprised 50.0% of the zooplankton mean biomass due to their larger size (Table 10). The cladoceran genus *Bosmina* represented the greatest percentage of biomass (44.9%), followed by the copepod genus *Epischura* (34.0%). The remaining biomass was mostly comprised of *Cyclops* (9.8%) and *Diaptomus* (5.0%).

The copepod *Diaptomus* was the largest zooplankter, having a mean length of 0.91 mm (Table 9). Of the remaining copepods, *Epischura* had a mean length of 0.87 mm, and *Cyclops* had a mean length of 0.71 mm. *Daphnia*, the largest cladoceran, had a mean length of 0.57 mm followed by *Holopedium* (0.49 mm), *Bosmina* (0.31 mm), and the “other cladocerans” made up of unidentified immature cladocerans that were too small to measure.

DISCUSSION

SMOLT ASSESSMENT

This was the fifth consecutive year of the Afognak Lake smolt assessment study in which the same methods and materials were used. The data collected from the prior years indicated that average trap efficiencies were consistent from year to year, despite the fact that seasonal conditions, water levels, and field personnel operating the system varied annually. The annual mean trap efficiencies for the preceding four-year study were within 5.0% of each other (2003: 19.9% 2004: 18.6%; 2005: 14.9%; 2006: 19.5%; Appendix A1). In 2007, the total trap efficiency (19.9%) also fell within the range of prior years. These results suggest that reliable and comparable estimates of annual smolt production have been made each year of the study.

Confidence in the annual estimates is strengthened through an alternative and distinctly different life history method of estimating smolt numbers by using known escapement and fecundity data with egg-to-smolt survival rates as reported from other sockeye salmon lake systems. The life history method derived theoretical production estimates, which is the number of smolt expected to have emigrated from Afognak Lake, from 2003-2007 and estimated an average of 432,962 smolt per year (Figure 6). The life history production estimates were compared to the average trap catch estimates of 407,126 smolt (total 2,035,630) for the same five year period. The difference between the two methods was considered minor (6%). Our confidence in the two estimates is strengthened by the fact that these two methods of estimating production are distinctly different and yet produced very similar results. Moreover, the theoretical estimation method utilized survival (egg-to-fry and fry-to-smolt) estimates that were based on multi-year data sets obtained from 18 different sockeye salmon systems (Koenings and Kyle 1997). These results suggest that the rearing conditions in Afognak Lake are supporting the current escapement level and that freshwater survival and production is at expected levels.

The 2007 emigration was dominated by age-1. smolt (86.2%) while the age-2. smolt made up the remaining (13.8%) portion of the emigration. Typically, systems producing large proportions of age-1. smolt have favorable freshwater rearing conditions. An increased proportion of older smolt often indicates decreased food availability due to either declining lake productivity or increasing numbers of juvenile salmon (Barnaby 1944; Burgner 1964; Foerster 1968; Koenings et al. 1993 Krokhin 1957). When the juvenile population begins to exceed the rearing capacity of a system, a greater proportion of the population must spend two or more years in freshwater before growing large enough to transform into smolt (Honnold and Schrof 2004). Based on the average dominance (81.8%) of age-1. smolt emigrating from Afognak Lake during 2003-2007, freshwater rearing capacity does not appear to have been exceeded and has been sufficient to support the juvenile population produced from recent escapements (Figure 7).

Mean size and condition of age-1. smolt sampled in 2007 (n=960; 2.6 g, 70.4 mm, 0.75 K) indicated a drop in comparison to recent year (2003-2006) averages of age-1. smolt (n=1,104; 3.7 g, 75.6 mm, 0.82 K; Appendix A2). In 2007, there was a late spring break up and a colder start to the summer, which may have led to a later development of phytoplankton resulting in later growth and development of zooplankton for foraging smolt. Smolt migrating out of Afognak Lake in the later portion of the sampling period were larger and had improved condition factors (age-1. smolt 0.80 K; Table 5).

Emigration timing of sockeye salmon smolt from Afognak Lake in 2007 was slightly later than the timing observed from 2003-2006 (Appendix A3). Smolt emigration typically begins in mid-May, peaks early to mid-June, and is over by early July. Smolt emigration in 2007 did not begin

until the end of May, but as in prior years, it peaked in mid-June and ended in early July. Documentation from other systems salmon (Barnaby 1944; Burgner 1964; Foerster 1968; Koenings et al. 1993 Krokhin 1957) indicates that older and larger smolt tend to migrate earlier. This later start of the emigration and smaller smolt size corroborates the theory that colder spring conditions led to later plankton development and slower smolt growth.

LIMNOLOGICAL ASSESSMENT

Most of the seasonal mean measurements of lake physical properties in 2007 were consistent with those from past years. Water temperatures were the exception with colder than the seasonal average readings from 1989 to 2006 (Appendix A4). With a mean depth of 8.6 m and a maximum depth of 23.0 m, Afognak Lake is considered a shallow lake that is easily influenced and mixed by wind and ice melt (Cole 1983). As a result of these mixing events, Afognak Lake is typically stratified into warm epilimnion and cool hypolimnion layers for only short periods of time each year. In 2007, a delay in lake turn-over and mixing was observed; the summer bottom temperature was nearly three degrees colder (9.5°C) than the average summer bottom temperature (12.5°C; Appendix A4). High dissolved oxygen levels recorded in 2007 were consistent with historical averages (Appendix A5). Light, euphotic volume and euphotic zone data recorded in 2007 were also similar to recent ten year averages (1987-2006; Appendix A6).

Historical nutrient and algal pigment concentrations have exhibited high levels of annual variation and irregular fluctuations, although notable differences were discernable between the eleven-year period in which the lake was fertilized (1990-2000) and the last seven years when the lake was not artificially fertilized (2001-2007; Appendix A7). The average TP, TKN, and NH_4^+ were higher during the fertilization years as compared to the non-fertilization years. During 2007, some surface water nutrient concentrations (TP: 3.6 $\mu\text{g L}^{-1}$ and TKN: 115 $\mu\text{g L}^{-1}$) were lower than the overall average concentration during the previous six-year (2001-2006) post-fertilization period (TP: 7.6 $\mu\text{g L}^{-1}$ and TKN: 135 $\mu\text{g L}^{-1}$), while the $\text{NO}_3^- + \text{NO}_2^-$ (56 $\mu\text{g L}^{-1}$) concentrations in 2007 were higher than the post fertilization year averages (41 $\mu\text{g L}^{-1}$). Seasonal average algal standing crop in 2007, as measured by chl *a* concentration (1.47 $\mu\text{g L}^{-1}$), was slightly less than the average concentration during the previous six-year post-fertilization period (1.60 $\mu\text{g L}^{-1}$) and slightly less than the average standing crop during the fertilization period (1.54 $\mu\text{g L}^{-1}$). The largest chl *a* fluctuations measured in Afognak Lake occurred during the fertilization years, during which concentrations ranged from 0.10 (in 1998) to 3.92 (in 1995) $\mu\text{g L}^{-1}$. Large fluctuations in chl *a* concentrations have been measured at other lakes on Afognak Island (Schrof and Honnold 2003).

Seasonal mean water chemistry has not varied a great deal, although average pH and alkalinity were both lower during the fertilization period (pH: 6.8; alkalinity: 9.8) than during the post-fertilization period (pH: 7.0; alkalinity: 10.3; Appendix A8). During 2007, average pH (6.8) was the same as the overall average for the fertilization period and slightly less than the previous six-year post-fertilization period. Average alkalinity for 2007 (10.9 mg L^{-1}) was greater than the overall averages for the fertilization and previous six-year post-fertilization periods.

During 2007, seasonal mean zooplankton density (84,151 individuals m^{-2}) and biomass (133 mg m^{-2}) estimates at Station 2 were much less than estimates from Station 1 (100,934 individuals m^{-2} and 120 mg m^{-2} ; Appendix A9). This was likely due to Station 2 being closer to the lake outlet. Lake water residence time is estimated to be only 0.4 years for Afognak Lake, so rapid lake flushing may remove zooplankton quicker than they can be replenished through reproduction

(Schrof and Honnold 2005; White et al. 1990). Rapid flushing may have also affected nutrient availability for phytoplankton, which could affect zooplankton production.

Since the zooplankton community serves as the primary forage base in lakes for juvenile sockeye salmon, total zooplankton density and biomass are often used as a measure to assess juvenile sockeye salmon production potential (Koenings et al. 1987). In evaluating these annual zooplankton estimates, the “other cladocerans” and “other copepods”, which are mostly composed of copepod larvae and unidentified immature cladoceran, were not included in the total estimates. During 2007, Station 1 weighted mean total zooplankton density (100,934 individuals m^{-2}) and biomass (120 $mg\ m^{-2}$) levels were slightly less than estimates for the pre-fertilization period (1987-1989: 134,747 no. m^{-2} and 194 $mg\ m^{-2}$) and very similar to the previous six-year post-fertilization period (2001-2006: 108,652 individuals and 139 $mg\ m^{-2}$; Appendix A9). Since juvenile sockeye salmon prefer cladocerans rather than copepods, cladoceran abundance is viewed as a better indicator of potential juvenile sockeye salmon production (Koenings et al. 1987; Kyle 1996).

The 2007 abundance of the cladoceran *Daphnia* at station 1 (3,386 individuals m^{-2} and 5 $mg\ m^{-2}$) was much greater than the overall average abundance during the pre-fertilization period (1,986 individuals and 3 $mg\ m^{-2}$) but was less than the previous six-year post-fertilization period (5,073 individuals and 7 $mg\ m^{-2}$; Appendix A9). The presence and abundance of *Daphnia*, a primary prey item for juvenile sockeye salmon, is considered an important indicator of lake forage activity (Honnold and Schrof 2001; Kyle 1996). Similar to *Daphnia*, the cladoceran *Holopedium* had a density (1,730 individuals m^{-2}) and biomass (3 $mg\ m^{-2}$) in 2007 that was just slightly less than the pre-fertilization years (1,716 individuals m^{-2} and 4 $mg\ m^{-2}$) and the post-fertilization years (2,589 individuals m^{-2} and 4 $mg\ m^{-2}$ Appendix A9). The abundance and the biomass of the cladoceran *Bosmina* (74,257 individuals m^{-2} and 66 $mg\ m^{-2}$) in 2007 was less than the average from the pre-fertilization years (104,823 individuals m^{-2} and 99 $mg\ m^{-2}$) but was more than the average from the post-fertilization years (66,443 individuals m^{-2} and 55 $mg\ m^{-2}$). Despite *Bosmina* comprising the vast majority of the cladocerans (89% of total Cladoceran biomass), it may not be a good indicator of available forage. Being about half the size of *Daphnia* and about two-thirds the size of *Holopedium*, *Bosmina* (0.31 mm) is a more difficult prey item for juvenile salmon to locate and eat due to its small size in the lake (Koenings and Kyle 1997).

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TABLES AND FIGURES

Table 1.–Afognak Lake sockeye salmon escapement, harvest, and total run estimates, 1978-2007.

Year	Escapement	Harvest			Total	Total Run
		Commercial ^a	Subsistence ^b	Sport ^c		
1978	52,701	3,414	1,632	524	5,570	58,271
1979	82,703	2,146	2,069	524	4,739	87,442
1980	93,861	28	3,352	524	3,904	97,765
1981	57,267	16,990	3,648	524	21,162	78,429
1982	123,055	21,622	3,883	524	26,029	149,084
1983	40,049	4,349	3,425	524	8,298	48,347
1984	94,463	6,130	3,121	524	9,775	104,238
1985	53,563	1,980	6,804	524	9,308	62,871
1986	48,328	2,585	3,450	524	6,559	54,887
1987	25,994	1,323	2,767	524	4,614	30,608
1988	39,012	14	2,350	524	2,888	41,900
1989	88,825	0	3,859	524	4,383	93,208
1990	90,666	22,149	4,469	524	27,142	117,808
1991	88,557	47,237	5,899	524	53,660	142,217
1992	77,260	2,196	4,638	600	7,434	84,694
1993	71,460	1,848	4,580	524	6,952	78,412
1994	80,570	17,362	3,329	524	21,215	101,785
1995	100,131	67,665	4,390	524	72,579	172,710
1996	101,718	106,141	11,023	258	117,422	219,140
1997	132,050	10,409	12,412	535	23,356	155,406
1998	66,869	26,060	4,690	718	31,468	98,337
1999	95,361	34,420	5,628	237	40,285	135,646
2000	54,064	14,124	7,572	364	22,060	76,124
2001	24,271	0	4,720	169	4,889	29,160
2002	19,520	0	1,279	41	1,320	20,840
2003	27,766	0	604	0	604	28,370
2004	15,181	0	567	10	577	15,758
2005	21,577	356	696	134	1,186	22,763
2006	22,933	6	451	36	493	23,426
2007	21,070	0	417	ND	417	21,487

^a Statistical fishing section 252-34 (Afognak Bay).

^b Data from ADF&G subsistence catch database 1978-2007.

^c Data from ADF&G Sport Fish Division statewide harvest survey (SWHS) for 1992, 1996-2006; SWHS data for other years did not have enough respondents to provide reliable estimates. Four years with reliable data were averaged and entered for years with no data.

Table 2.—Sockeye salmon smolt counts, number of samples collected, mark-recapture counts, and trap efficiency ratios from trapping at Afognak River, 2007.

Date	Catch Daily	Catch Cumulative	Dye Test Period Cumulative	AWL Sample Cumulative	Number Marked Releases ^a	Marked Recoveries Cumulative	Trap Efficiency (%)
10-May	1	1					
11-May	5	6					
12-May	0	6					
13-May	5	11					
14-May	2	13					
15-May	4	17					
16-May	7	24					
17-May	18	42					
18-May	7	49					
19-May	19	68					
20-May	33	101		20			
21-May	34	135		30			
22-May	117	252		70			
23-May	307	559		110			
24-May	145	704		110			
25-May	64	768		110			
26-May	41	809		110			
27-May	41	850		110			
28-May	73	923		130			
29-May	722	1,645		170			
30-May	729	2,374		210	415	51	
31-May	1,353	3,727		250		51	
1-Jun	1,853	5,580		250		51	
2-Jun	2,043	7,623		250		51	
3-Jun	1,357	8,980		290		51	
4-Jun	2,273	11,253		330		51	
5-Jun	3,197	14,450	14,450	370		51	12.5%
6-Jun	4,294	18,744		410	202	88	
7-Jun	2,352	21,096		450		111	
8-Jun	2,183	23,279		450		124	
9-Jun	2,946	26,225		450		124	
10-Jun	1,174	27,399		490		124	
11-Jun	3,251	30,650		530		124	
12-Jun	3,269	33,919	19,469	570		124	61.5%

-continued-

Table 2.--(page 2 of 2)

Date	Catch Daily	Catch Cumulative	Dye Test Period Cumulative	AWL Sample Cumulative	Number Marked Releases ^a	Marked Recoveries Cumulative	Trap Efficiency (%)
13-Jun	2,072	35,991		610	510	38	
14-Jun	3,213	39,204		650		77	
15-Jun	1,704	40,908		650		81	
16-Jun	2,156	43,064		650		82	
17-Jun	2,222	45,286		690		82	
18-Jun	987	46,273		730		82	
19-Jun	1,674	47,947		770		82	
20-Jun	1,253	49,200	13,209	810		82	16.2%
21-Jun	1,399	50,599		850	541	81	
22-Jun	1,099	51,698		850		107	
23-Jun	1,072	52,770		890		108	
24-Jun	684	53,454		890		108	
25-Jun	419	53,873		930		108	
26-Jun	399	54,272		970		108	
27-Jun	144	54,416	5,216	1,010		108	20.1%
28-Jun	349	54,765		1050	401	36	
29-Jun	294	55,059		1050		43	
30-Jun	140	55,199		1,090		44	
1-Jul	55	55,254				44	
2-Jul	8	55,262				44	
3-Jul	34	55,296				44	
4-Jul	19	55,315	899			44	11.2%
5-Jul	Trap Pulled				Average Trap Efficiency=		19.9%

^a Adjusted number released using the delayed mortality formulation.

Table 3.–Population estimate of the sockeye salmon smolt emigration from Afognak Lake, 2007.

Stratum (h)	Beginning Date	Ending Date	uh Unmarked	Mh Released	mh Recovered	Uh Estimate	var(Uh) Variance	95% Confidence Interval	
								lower	upper
1	5/10	6/5	14,450	415	51	115,690	2.22E+08	86,501	144,879
2	6/6	6/12	19,469	202	124	31,680	3.09E+06	28,235	35,125
3	6/13	6/20	15,281	510	82	94,135	8.88E+07	75,660	112,609
4	6/21	6/27	5,216	541	108	25,914	4.98E+06	21,541	30,288
5	6/28	7/4	899	401	44	8,031	1.31E+06	5,790	10,272
Total						275,450	3.20E+08	240,388	310,512
						SE=	17,889		

Table 4.—The Afognak Lake sockeye salmon smolt emigration estimate based on percents by age class and dye test period, 2007.

Stratum		Age			Total
		1.	2.	3.	
1 (5/10-6/5)	Number	79,916	35,774	0	115,690
	Percent	69.1%	30.9%	0.0%	100.0%
2 (6/6-6/12)	Number	29,779	1,901	0	31,680
	Percent	94.0%	6.0%	0.0%	100.0%
3 (6/13-6/20)	Number	93,742	392	0	94,135
	Percent	99.6%	0.4%	0.0%	100.0%
4 (6/21-6/27)	Number	25,914	0	0	25,914
	Percent	100.0%	0.0%	0.0%	100.0%
5 (6/28-7/4)	Number	8,031	0	0	8,031
	Percent	100.0%	0.0%	0.0%	100.0%
Total	Number	237,383	38,067	0	275,450
	Percent	86.2%	13.8%	0.0%	100.0%

Table 5.—Length, weight, and condition of sockeye salmon smolt from the Afognak River, 2007.

Stratum	Dates	Sample Size	Weight (mm)		Length (g)		Condition	
			Mean	Standard Error	Mean	Standard Error	Mean	Standard Error
Age 1.								
1	5/10-6/5	213	2.4	0.02	69.1	0.24	0.73	0.006
2	6/6-6/12	188	2.3	0.02	69.1	0.17	0.70	0.005
3	6/13-6/20	239	2.5	0.02	69.6	0.14	0.74	0.004
4	6/21-6/27	160	2.9	0.03	71.2	0.20	0.79	0.005
5	6/28-7/4	160	3.3	0.04	74.0	0.25	0.80	0.005
Totals		960	2.6	0.02	70.4	0.11	0.75	0.002
Age 2.								
1	5/10-6/5	116	3.4	0.08	77.1	0.46	0.74	0.01
2	6/6-6/12	12	2.6	0.11	71.3	0.79	0.72	0.025
3	6/13-6/20	1	3.2	0.00	73.0	0.00	0.82	0
4	6/21-6/27	0	—	—	—	—	—	—
5	6/28-7/4	0	—	—	—	—	—	—
Totals		129	3.4	0.07	76.5	0.44	0.74	0.009

Table 6.–Afognak Lake sockeye salmon theoretical production of eggs, emergent fry, and smolt by age from brood years 2004 and 2005 and predicted smolt emigration in 2007.

Parameter	Production		Brood Year		2007 Estimate Age 1. and 2. smolt
	Assumption		2004	2005	
Escapement			15,181	21,577	
Females spawning	1:1 sex ratio		7,591	10,789	
Deposited Eggs	2,112 per female ^a		16,031,136	22,785,312	
Emergent Fry	7% egg-to-fry survival ^b		1,122,180	1,594,972	
Smolt	21% fry-to-smolt survival ^c		235,658	334,944	
Emigration in 2007	86.2% age-1., 13.8% age-2. ^d		32,521	288,722	321,243

^a Average fecundity as reported from Pillar Creek Hatchery (2003-2006)

^b Averages from Drucker (1970), Bradford (1995) and Koenings and Kyle (1997)

^c Koenings and Kyle (1997)

^d Based on 2007 smolt age class estimates (Table 4)

Table 7.—General water chemistry and algal pigment concentrations at 1 m water depth, station 1, Afognak Lake 2007.

Date	pH (units)	Alkalinity (mg L ⁻¹)	Silicon (µg L ⁻¹)	Chlorophyll <i>a</i> (µg L ⁻¹)
18-May	6.70	10.0	ND	0.96
26-Jun	6.63	10.0	ND	1.28
2-Aug	6.89	10.3	ND	2.24
29-Aug	6.79	11.8	ND	1.60
25-Sep	6.84	12.5	ND	1.28
Average	6.77	10.9	ND	1.47
SD	0.11	1.2	ND	0.49

Table 8.–Seasonal phosphorus and nitrogen concentrations at 1 m water depth, station 1, Afognak Lake, 2007.

Date	Total filterable-P ($\mu\text{g L}^{-1}$)	Filterable reactive-P ($\mu\text{g L}^{-1}$)	Total-P ($\mu\text{g L}^{-1}$)	Ammonia ($\mu\text{g L}^{-1}$)	Total Kjeldahl Nitrogen ($\mu\text{g L}^{-1}$)	Nitrate + Nitrite ($\mu\text{g L}^{-1}$)	Total Nitrogen ($\mu\text{g L}^{-1}$)	TN:TP ratio
18-May	1.3	1.9	3.4	5.3	178.0	124.5	302.5	197.0
26-Jun	1.6	0.7	3.3	5.2	99.0	75.1	174.1	116.8
2-Aug	1.0	0.8	3.7	4.6	102.0	21.3	123.3	73.8
29-Aug	0.6	0.4	4.3	6.3	110.0	24.5	134.5	69.3
25-Sep	1.1	1.9	3.2	6.5	86.0	32.3	118.3	81.9
Average	1.1	1.1	3.6	5.6	115.0	55.5	170.5	107.7
SD	0.4	0.7	0.4	0.8	36.3	44.2	77.0	53.3

Table 9.—Weighted mean zooplankton density, biomass, and size by station from Afognak Lake, 2007.

Station	<i>n</i>		<i>Epischura</i>	<i>Diaptomus</i>	<i>Cyclops</i>	Other Copepods	<i>Bosmina</i>	<i>Daphnia</i>	<i>Holopedium</i>	Other Cladocerans	Total Copepods	Total Cladocerans	Total all zooplankton
1	5	density (no. m ⁻²)	10,913	2,930	7,718	18,631	74,257	3,386	1,730	3,397	40,192	82,770	122,962
		%	8.9%	2.4%	6.3%	15.2%	60.4%	2.8%	1.4%	2.8%	32.7%	67.3%	100.0%
		biomass (mg m ⁻²)	24.0	8.9	13.0	2.3	66.0	5.0	3.0	0.1	48.2	74.1	122.3
		%	19.6%	7.3%	10.6%	1.9%	54.0%	4.1%	2.5%	0.1%	59.8%	60.6%	100.0%
		size (mm)	0.78	0.88	0.70	0.58	0.31	0.58	0.47	0.41			
2	5	density (no. m ⁻²)	16,269	1,141	6,693	16,174	57,065	934	2,049	1,672	40,277	61,720	101,997
		%	16.0%	1.1%	6.6%	15.9%	55.9%	0.9%	2.0%	1.6%	39.5%	60.5%	100.0%
		biomass (mg m ⁻²)	63.0	4.0	12.0	0.7	49.0	1.0	4.0	na	79.7	54.0	133.7
		%	47.1%	3.0%	9.0%	0.5%	36.7%	0.7%	3.0%	na	59.6%	40.4%	100.0%
		size (mm)	0.95	0.93	0.71	0.58	0.31	0.55	0.50				
1 & 2 Averaged		density (no. m ⁻²)	13,591	2,036	7,206	17,403	65,661	2,160	1,890	2,535	40,235	72,245	112,480
		%	12.1%	1.8%	6.4%	15.5%	58.4%	1.9%	1.7%	2.3%	35.8%	64.2%	100.0%
		biomass (mg m ⁻²)	43.5	6.5	12.5	1.5	57.5	3.0	3.5	na	63.9	64.0	127.9
		%	34.0%	5.0%	9.8%	1.2%	44.9%	2.3%	2.7%	na	50.0%	50.0%	100.0%
		size (mm)	0.87	0.91	0.71	0.58	0.31	0.57	0.49				

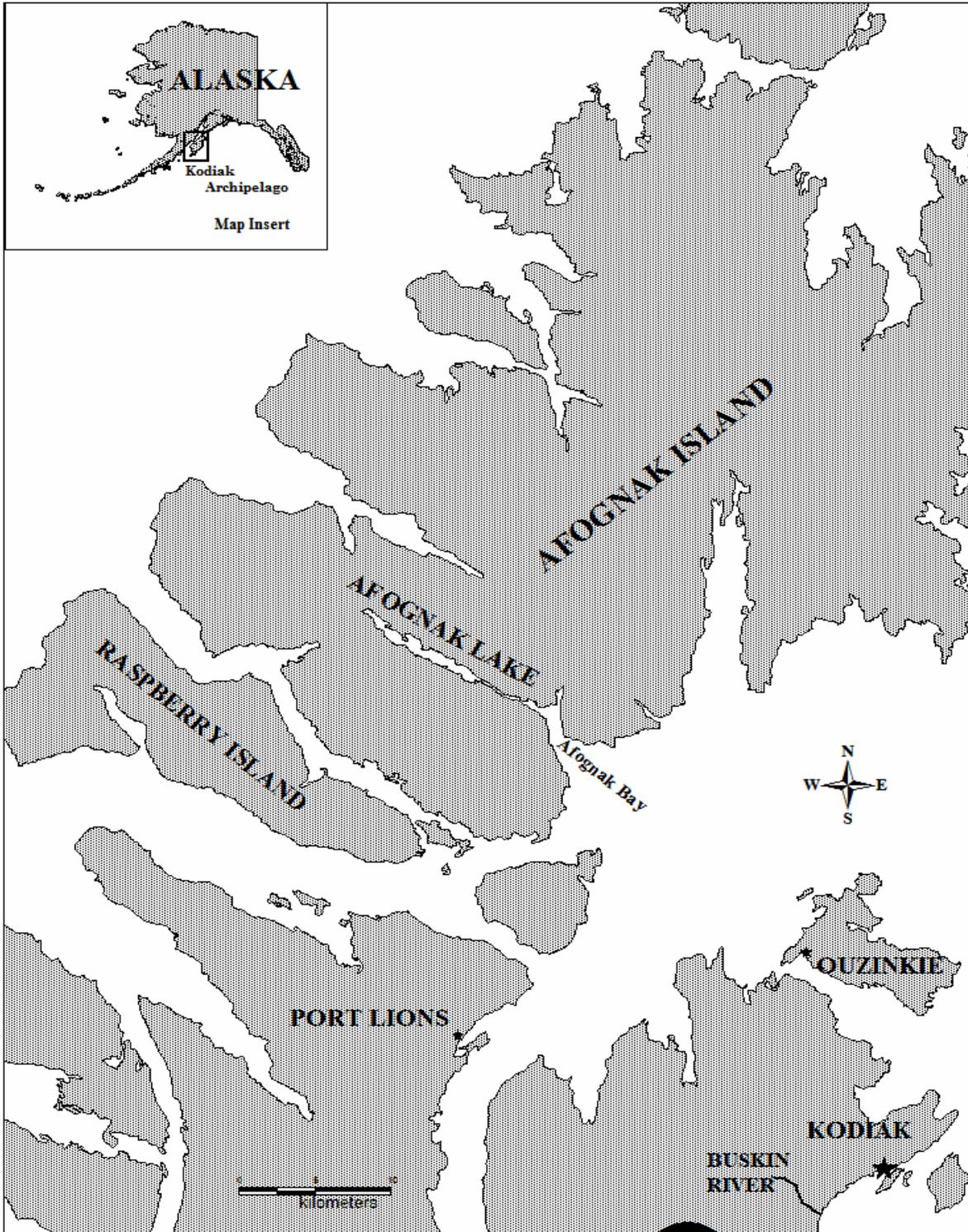


Figure 1.—This map displays the location of Kodiak City, and the villages of Port Lions, and Ouzinkie and their proximity to the Afognak Lake drainage on Afognak Island.

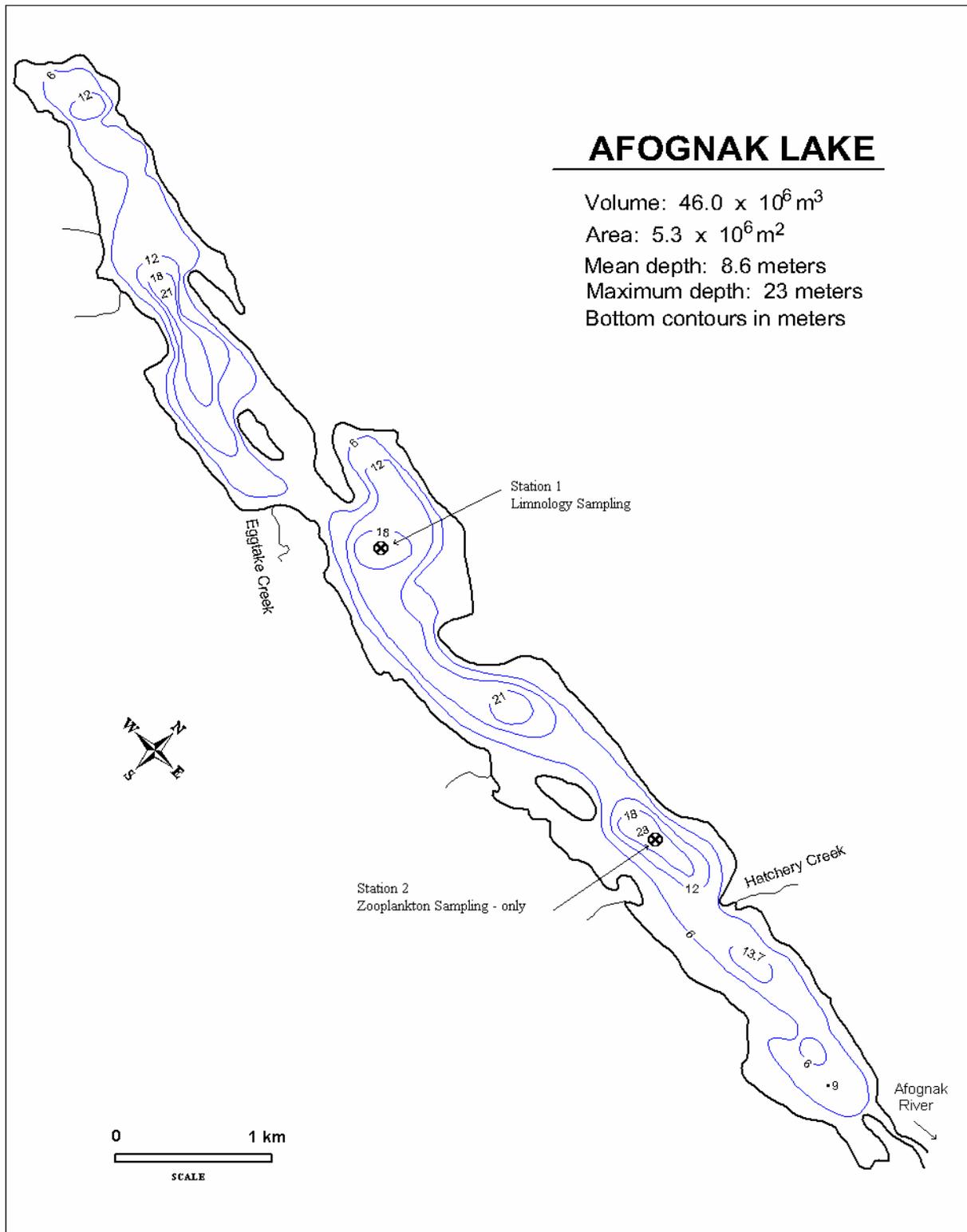


Figure 2.–Bathymetric map showing the two limnology and zooplankton stations on Afognak Lake.



Figure 3.—The smolt trapping system set up in the Afognak River, 2007.

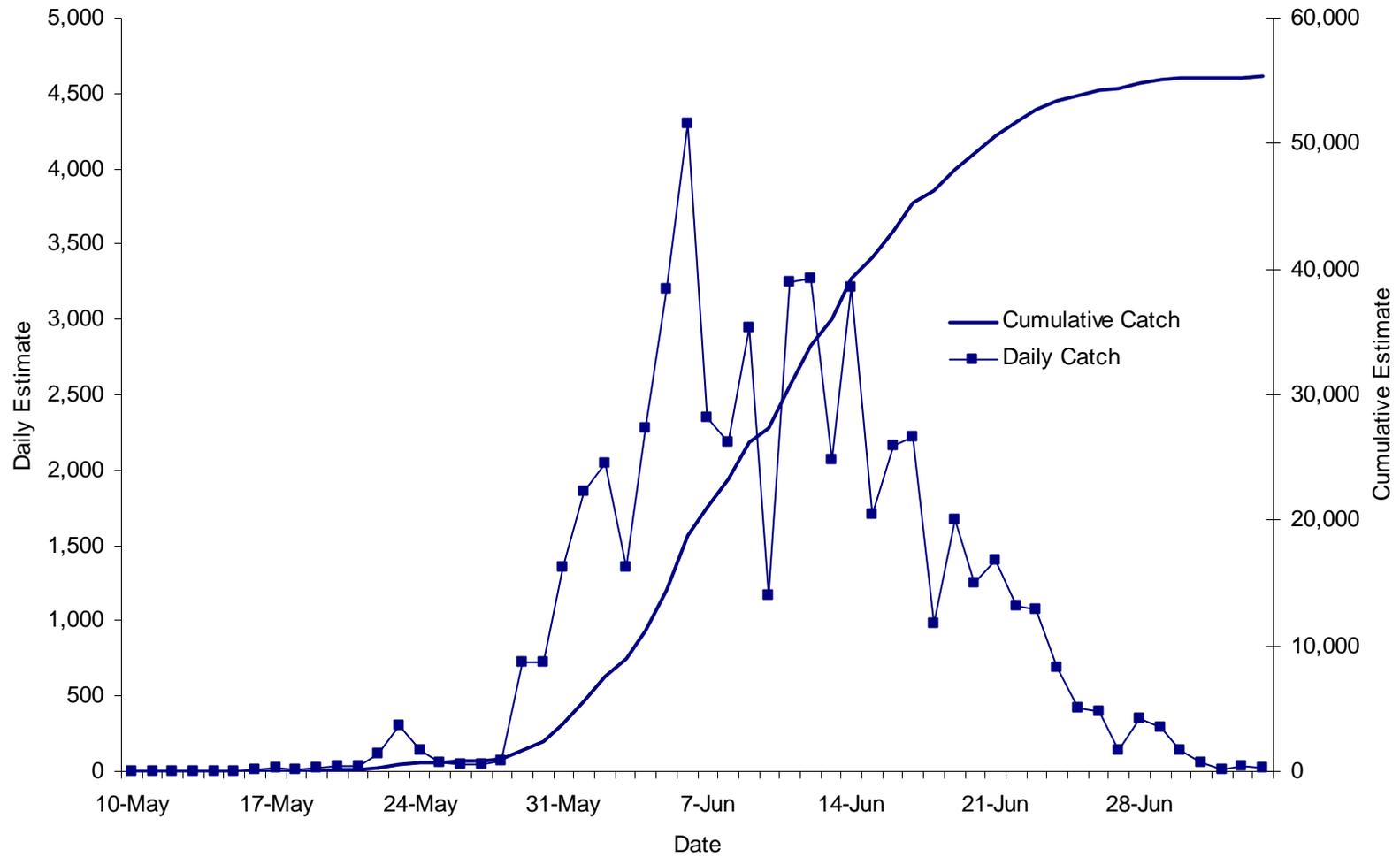


Figure 4.—Daily and cumulative sockeye salmon smolt trap catch estimates from 10 May to 4 July in the Afognak River, 2007.

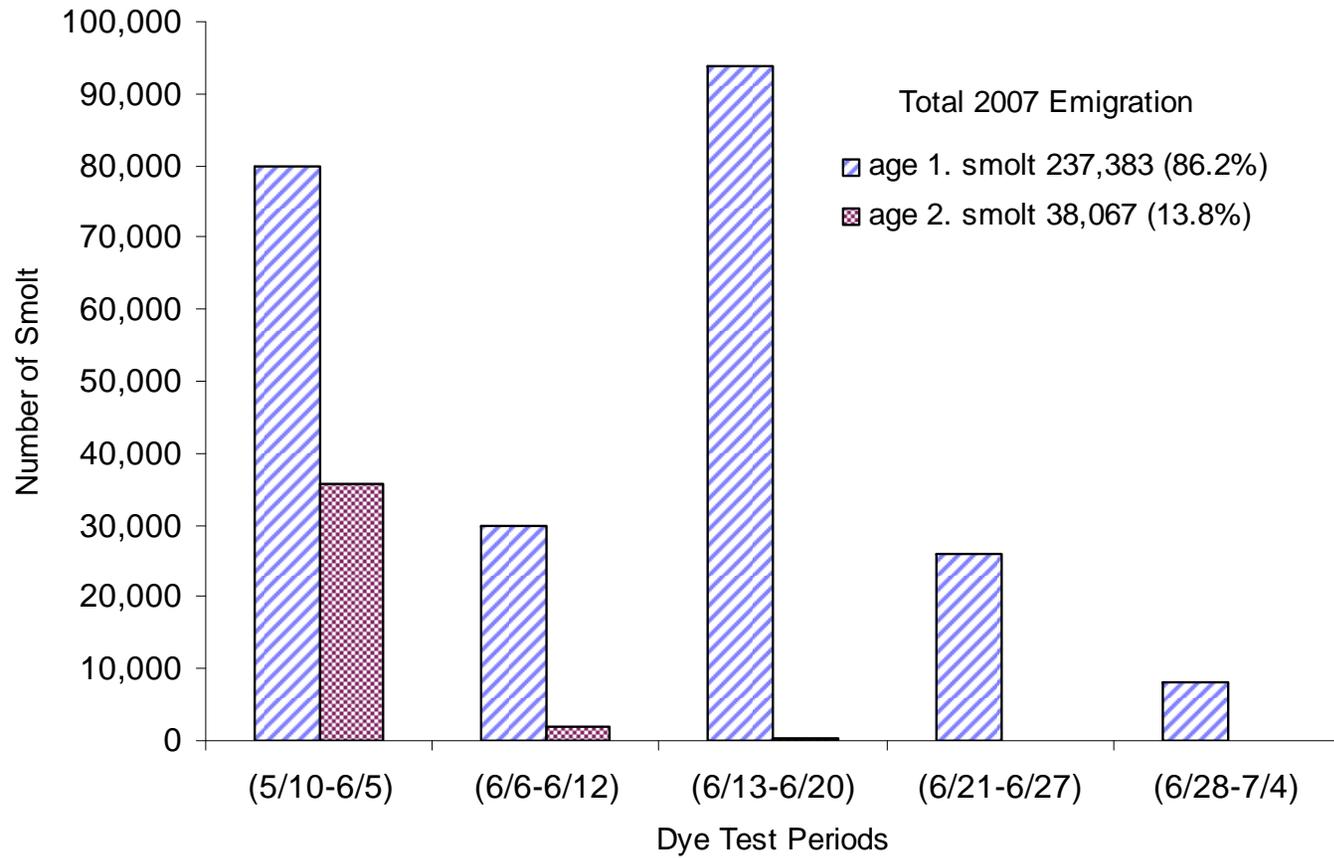


Figure 5.—Afognak Lake sockeye salmon smolt emigration by age class and dye test period, 2007.

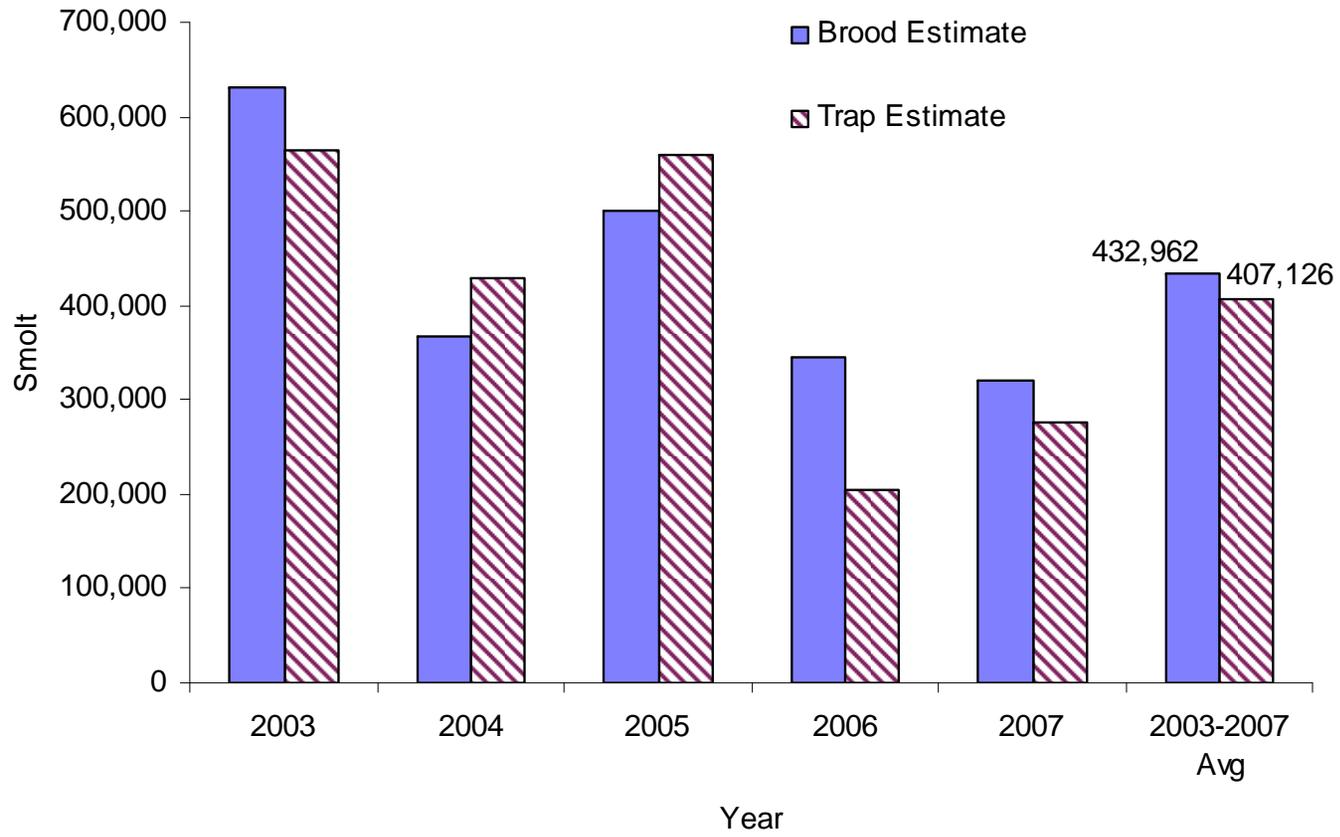


Figure 6.—Afognak Lake emigration estimates from trap catches and theoretical emigration estimates based on brood year escapements, 2003-2007.

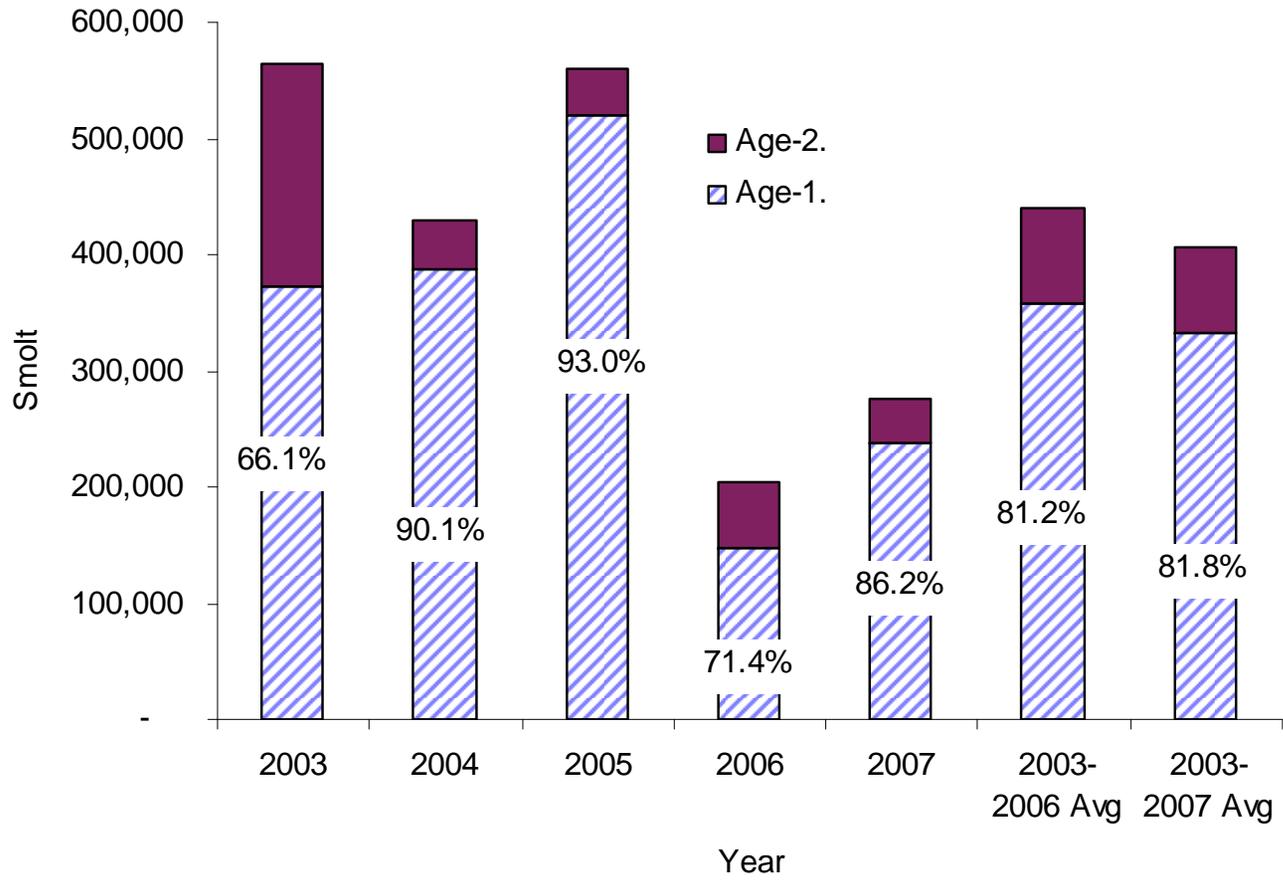


Figure 7.—Sockeye salmon smolt emigration by age from Afognak Lake, 2003-2007.

**APPENDIX A. SUPPORTING HISTORICAL
INFORMATION**

Appendix A1.—Population estimates of the sockeye salmon emigrations from Afognak Lake 2003-2007.

Stratum (h)	Starting Date	Ending Date	Catch (u_h)	Released (M_h)	Recaptured (m_h)	Avg. Trap Efficiency (%)	Estimate (U_h)	Variance var (U_h)	95% Confidence Interval	
									lower	upper
2003										
1	5/12	5/19	1,387	239	5	2.1%	55,480	4.31E+08	14,809	96,151
2	5/20	5/25	2,912	239	5	2.1%	116,480	1.89E+09	31,188	201,772
3	5/26	5/31	11,966	706	161	22.8%	52,222	1.31E+07	45,136	59,308
4	6/1	6/7	31,358	638	133	20.8%	149,536	1.31E+08	127,063	172,008
5	6/8	6/10	11,153	686	257	37.5%	29,698	2.18E+06	26,807	32,589
6	6/11	6/18	18,696	679	103	15.2%	122,243	1.21E+08	100,663	143,823
7	6/19	6/26	4,762	506	79	15.6%	30,179	9.63E+06	24,097	36,261
8	6/27	7/3	736	218	17	7.8%	8,955	3.97E+06	5,050	12,859
Total			82,970	3,911	760	19.9%	564,793	2.61E+09	374,814	754,772
							SE=	51,047		
2004										
1	5/11	5/26	24,278	525	56	10.7%	224,039	7.73E+08	169,530	278,548
2	5/27	6/3	17,727	547	96	17.6%	100,148	8.47E+07	82,111	118,186
3	6/4	6/11	16,658	700	211	30.1%	55,081	1.01E+07	48,864	61,299
4	6/12	6/19	5,086	613	119	19.4%	26,023	4.61E+06	21,815	30,231
5	6/20	7/3	3,779	581	88	15.1%	24,712	5.88E+06	19,958	29,466
Total			67,528	2,966	570	18.6%	430,004	8.79E+08	371,905	488,104
							SE=	29,643		

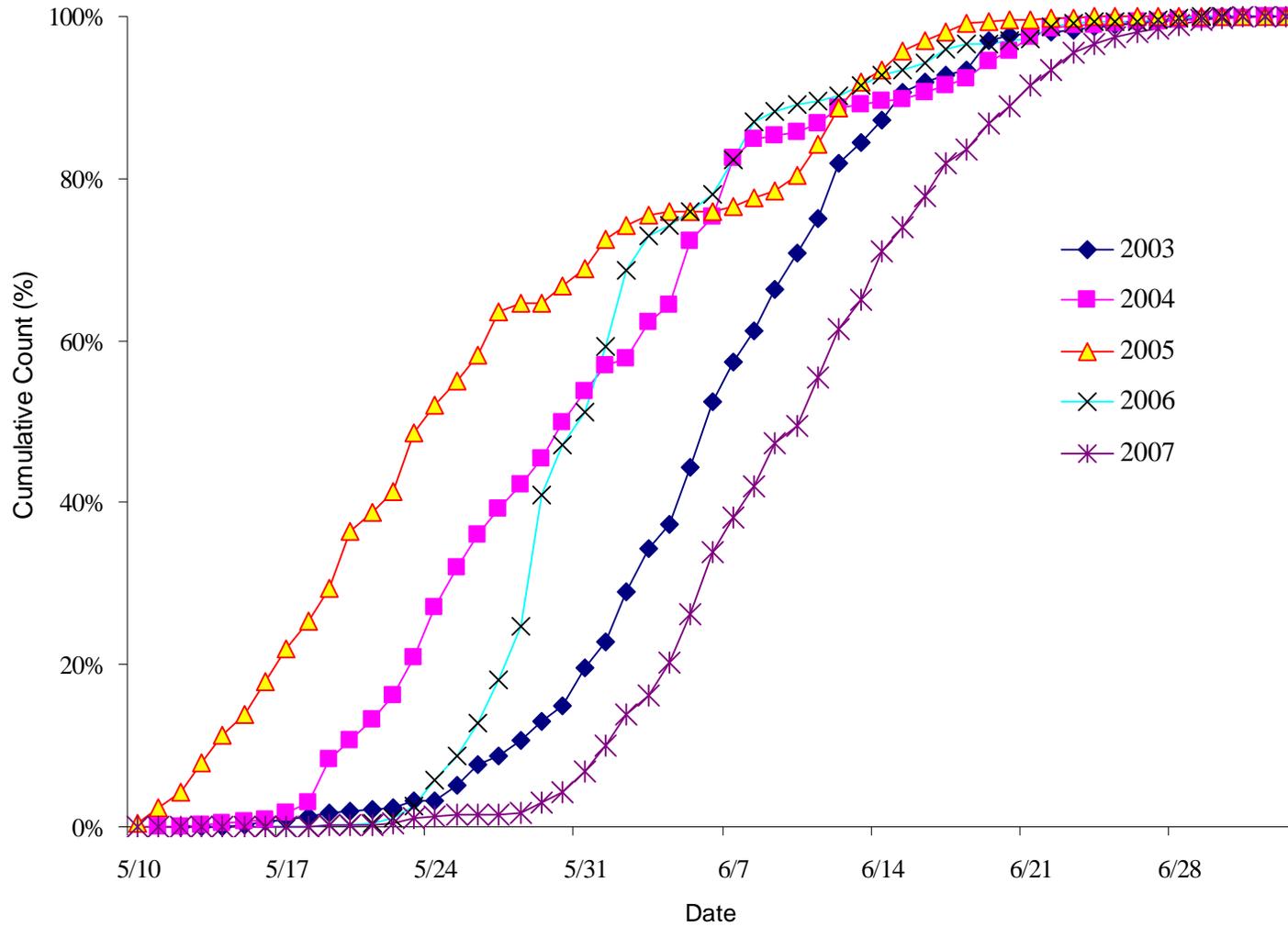
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Stratum (h)	Starting Date	Ending Date	Catch (u_h)	Released (M_h)	Recaptured (m_h)	Avg.Trap Efficiency (%)	Estimate (U_h)	Variance var (U_h)	95% Confidence Interval	
									lower	upper
2005										
1	5/10	5/21	27,226	489	70	14.3%	184,879	4.05E+08	145,443	224,314
2	5/22	5/26	13,627	518	43	8.3%	155,259	4.89E+08	111,932	198,587
3	5/27	6/5	15,210	482	44	9.1%	158,499	4.94E+08	114,948	202,050
4	6/6	6/27	17,634	368	103	28.0%	61,593	2.58E+07	51,640	71,546
Total			73,697	1,857	260	14.9%	560,230	1.41E+09	486,554	633,906
							SE=	37,590		
2006										
1	5/16	6/1	25,983	312	73	23.6%	110,017	1.24E+08	88,224	131,809
2	6/2	6/6	8,199	515	98	19.2%	42,726	1.49E+07	35,153	50,299
3	6/7	6/16	7,108	485	95	19.8%	35,975	1.09E+07	29,519	42,432
4	6/17	6/29	2,534	492	75	15.4%	16,435	3.06E+06	13,009	19,861
Total			43,824	1,804	341	19.5%	205,153	1.52E+08	180,952	229,353
							SE=	12,347		
2007										
1	5/10	6/5	14,450	415	51	12.5%	115,690	2.22E+08	86,501	144,879
2	6/6	6/12	19,469	202	124	61.5%	31,680	3.09E+06	28,235	35,125
3	6/13	6/20	15,281	510	82	16.2%	94,135	8.88E+07	75,660	112,609
4	6/21	6/27	5,216	541	108	20.1%	25,914	4.98E+06	21,541	30,288
5	6/28	7/4	899	401	44	11.2%	8,031	1.31E+06	5,790	10,272
Total			55,315	2,070	409	19.9%	275,450	3.20E+08	240,388	310,512
							SE=	17,889		

Appendix A2.—Mean weight, length, and condition coefficient by age for sockeye salmon smolt sampled at Afognak Lake, 1987-2001, and 2003-2007.

Year	Sampling Period	Age-1			Age-2				
		n	Weight (g)	Length (mm)	Condition (K)	n	Weight (g)	Length (mm)	Condition (K)
1987	8-Jun	36	3.6	74.9	0.85	186	3.6	79.3	0.86
1988	15-Jun	202	4.1	77.9	0.90	0			
1989	15-Jun	208	4.1	76.8	0.91	2	5.2	78.0	1.10
1990	May 23-June 24	544	2.5	68.8	0.76	21	3.4	77.3	0.73
1991	May 13-June 26	1,895	3.1	72.9	0.78	176	3.9	78.3	0.81
1992	June 7-20	268	3.8	77.0	0.82	37	3.8	76.9	0.83
1993	May 24-30	274	3.0	72.7	0.78	21	3.3	74.8	0.79
1994	May 17-23	138	3.0	72.0	0.81	142	4.7	84.3	0.79
1995	May 31-June 13	394	2.8	69.4	0.84	5	3.6	78.8	0.74
1996	June 5-11	54	4.6	80.9	0.87	339	4.8	81.6	0.88
1997	May 24-30	76	4.3	81.7	0.78	122	4.4	82.1	0.79
1998	May 24-30	116	2.6	66.4	0.82	46	6.6	88.0	0.90
1999	May 31-June 6	96	2.8	74.6	0.66	98	2.1	66.6	0.69
2000	May 31-June 13	84	4.9	81.5	0.89	100	5.6	85.3	0.89
2001	June 11-13	44	7.0	90.1	0.93	17	5.8	85.6	0.92
2003	May 12-July 3	1,031	4.2	79.1	0.82	383	4.2	81.4	0.77
2004	May 11-July 3	1,370	3.6	75.7	0.80	81	3.6	78.7	0.74
2005	May 10-June 27	1,248	3.9	76.8	0.84	65	4.2	81.3	0.77
2006	May 16-June 29	765	3.0	70.8	0.83	202	3.8	79.6	0.75
2007	May 21 - July 2	960	2.6	70.4	0.75	129	3.4	76.5	0.74
2003-2006 Avg		1,104	3.7	75.6	0.82	183	4.0	80.3	0.76
2003-2007 Avg		1,075	3.5	74.6	0.81	172	3.8	79.5	0.75

Appendix A3.-Sockeye salmon smolt emigration timing from Afognak Lake, 2003-2007.



Appendix A4.—Temperatures (°C) measured at the 1-meter and near bottom strata in the Spring (May-June), summer (July-August), and fall (September-October) for Afognak Lake 1989-2007.

Year	Spring		Summer		Fall	
	Surface	Bottom	Surface	Bottom	Surface	Bottom
1989	7.8	7.0	16.3	12.8	15.3	13.6
1990	9.4	8.3	14.8	13.6	11.9	11.4
1991	6.2	5.7	15.1	12.5	12.4	12.1
1992	10.0	8.9	15.5	13.9	11.1	11.0
1993	11.9	10.4	17.6	14.5	13.5	12.6
1994	10.8	8.8	15.5	13.5	10.2	9.7
1995	8.8	7.3	15.2	12.8	12.5	11.9
1996	11.5	9.7	15.2	13.9	11.1	10.5
1997	10.3	7.5	17.6	10.6	14.1	12.4
1998	7.9	7.7	14.3	13.0	11.8	11.6
1999	7.0	6.2	15.1	11.4	10.4	10.1
2000	9.7	8.7	15.0	13.1	10.1	10.0
2001	9.1	7.0	17.1	10.2	12.9	12.5
2002	10.0	7.8	16.0	10.8	9.3	9.2
2003	9.7	5.5	18.3	12.9	11.5	11.3
2004	9.2	8.2	15.1	11.7	13.1	12.9
2005	11.8	9.5	18.1	13.5	13.6	13.5
2006	9.2	8.0	15.8	12.5	12.6	12.5
2007	9.2	6.7	15.4	9.5	12.4	12.3
Avg 1989-2006	9.4	7.9	16.0	12.6	12.1	11.6
Avg 1989-2007	9.4	7.8	15.9	12.5	12.1	11.6

Appendix A5.–Dissolved oxygen concentrations (mg L⁻¹) measured at the 1-meter and near bottom strata in the Spring (May-June), summer (July-August), and fall (September-October) for Afognak Lake 1989-2007.

Year	Spring		Summer		Fall	
	Surface	Bottom	Surface	Bottom	Surface	Bottom
1989	11.7	11.2	10.3	9.2	13.1	10.3
1990	14.0	11.8	9.5	8.6	9.6	8.9
1991	12.6	11.1	10.9	8.2	10.5	9.4
1992	11.5	10.8	10.1	8.7	10.8	10.8
1993	10.9	9.8	9.5	7.5	10.5	10.1
1994	11.0	9.8	10.0	8.1	11.3	10.9
1995	11.4	11.3	10.0	8.4	10.5	9.8
1996	10.9	10.5	10.0	7.7	11.2	11.1
1997	10.5	10.7	9.0	4.6	10.2	7.6
1998	11.8	11.7	10.2	6.1	10.2	10.0
1999	11.9	11.5	9.6	6.2	10.9	10.4
2000	11.0	9.1	9.7	6.8	10.5	10.1
2001	9.7	9.6	9.3	4.7	9.0	8.1
2002	10.8	9.3	9.8	0.1	10.5	10.1
2003	12.0	11.1	9.2	5.5	18.0	10.3
2004	12.9	11.2	11.5	8.1	10.5	6.4
2005	10.8	10.2	9.5	5.1	9.5	8.7
2006	10.9	10.0	9.8	8.3	10.5	10.0
2007	11.4	10.8	9.2	6.6	10.6	9.9
Avg 1989-2006	11.4	10.6	9.9	6.8	10.9	9.6
Avg 1989-2007	11.4	10.6	9.8	6.8	10.9	9.6

Appendix A6.—Average euphotic zone depth (EZD), light extinction coefficient (K_d), Secchi disk transparency, and euphotic volume (EV) from stations 1 and 2 for Afognak Lake, 1987-2007.

	EZD		K _d		Secchi Disk		EV	
	(m)	SD	(m ⁻¹)	SD	(m)	SD	(10 ⁶ m ³)	SD
1990	7.47	2.46	-2.01	0.53	3.6	0.6	39.60	13.02
1991	8.36	2.40	-2.25	0.68	2.7	0.5	44.32	12.75
1992	9.39	2.79	-2.28	0.35	2.8	0.9	49.77	14.77
1993	9.27	2.23	-2.09	0.52	3.5	0.5	49.14	11.81
1994	7.73	1.45	-1.86	0.33	3.4	0.4	40.97	7.67
1995	7.56	1.18	-1.79	0.27	2.5	0.6	40.08	6.23
1996	8.19	1.53	-1.92	0.37	3.5	0.4	43.41	8.13
1997	6.15	1.75	-1.68	0.59	3.2	0.7	32.61	9.27
1998	7.64	0.82	-1.76	0.25	3.8	1.2	40.50	4.36
1999	9.12	2.67	-1.82	0.35	2.9	0.6	48.36	14.14
2000	9.93	1.65	-2.28	0.39	3.4	0.6	52.62	8.76
2001	10.87	3.24	-2.24	0.40	4.0	1.1	57.61	17.17
2002	10.15	0.69	-2.43	0.17	4.3	0.5	53.80	3.66
2003	9.91	1.11	-2.36	0.25	4.5	0.2	52.51	5.87
2004	10.27	2.57	-2.32	0.31	4.0	0.3	54.42	13.60
2005	9.77	0.64	-2.28	0.20	4.7	0.6	51.77	3.37
2006	9.18	1.05	-2.16	0.36	4.0	0.7	48.67	5.54
2007	9.36	1.27	-2.05	0.36	4.1	0.7	49.61	6.73
Avg 1990-2006	8.88	1.78	-2.09	0.37	3.58	0.63	47.07	9.42
Avg 1990-2007	8.91	1.75	-2.09	0.37	3.61	0.63	47.21	9.27

Appendix A7.—Summary of seasonal mean nutrient and algal pigment concentrations by station and depth for Afognak Lake, 1990-2007.

Year	Station	Depth (m)	Total Phosphorus		Total filterable-P		Filterable reactive-P		Total Kjeldahl Nitrogen		Ammonia		Nitrate +Nitrate		Reactive Silicon		Organic Carbon		Chlorophyll <i>a</i>		Phaeophytin <i>a</i>	
			(ug L ⁻¹)	SD	(ug L ⁻¹)	SD	(ug L ⁻¹)	SD	(ug L ⁻¹)	SD	(ug L ⁻¹)	SD	(ug L ⁻¹)	SD	(ug L ⁻¹)	SD	(ug L ⁻¹)	SD	(ug L ⁻¹)	SD	(ug L ⁻¹)	SD
1990	1	1	4.5	1.5	2.9	4.2	3.7	1.7	128	16.5	8	3.0	40	29.1	3250	247.5	145	13.0	0.34	0.19	0.17	0.03
	1	16	5.1	2.3	1.3	1.3	2.8	1.1	118	22.7	10	4.2	65	29.1	3390	154.5	144	30.6	0.21	0.03	0.28	0.07
1991	1	1	5.0	2.8	3.2	0.6	2.3	0.4	151	22.6	11	1.8	57	21.3	2865	108.6	ND	ND	0.31	0.21	0.27	0.07
	1	14	4.6	1.5	6.0	3.5	4.5	3.2	138	12.3	14	5.0	70	23.2	2966	156.3	ND	ND	0.22	0.14	0.22	0.08
1992	1	1	3.8	0.5	4.1	2.5	3.1	2.4	135	13.9	3	1.7	62	26.1	3163	158.9	199	64.1	0.44	0.29	0.28	0.13
	1	24	3.9	1.7	4.0	3.2	2.6	1.7	127	12.8	10	4.1	93	23.1	3182	198.0	163	52.9	0.31	0.25	0.28	0.12
1993	1	1	4.5	0.8	3.7	1.3	2.8	0.5	148	18.5	5	2.2	49	30.4	3132	220.6	147	53.3	1.01	0.31	0.36	0.03
	1	25	4.9	1.3	8.5	11.7	6.8	9.9	136	17.3	19	10.1	98	31.7	3380	244.0	121	47.5	0.52	0.21	0.45	0.14
1994	1	1	5.7	0.7	4.5	3.3	3.6	2.3	160	23.8	3	1.7	40	21.4	2843	122.4	114	33.0	0.56	0.26	0.28	0.08
	1	2	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	0.56	0.34	0.34	0.10
	1	26	5.3	1.1	4.8	3.9	4.2	3.2	160	17.7	15	9.7	74	23.8	3177	285.5	128	52.1	0.36	0.21	0.27	0.09
1995	1	1	8.7	2.7	3.0	1.5	2.0	1.1	168	21.6	9	14.1	66	22.1	1873	735.0	ND	ND	3.92	2.44	1.13	0.62
	1	17	8.1	2.0	1.9	1.1	1.1	0.4	187	47.1	35	44.3	45	35.0	2046	618.4	ND	ND	3.13	1.75	1.10	0.54
	2	1	7.4	2.1	2.1	1.2	1.7	1.0	169	31.0	9	14.0	54	33.2	1942	753.9	ND	ND	4.20	2.90	1.05	0.65
	2	11	7.2	1.7	2.2	2.0	1.6	1.1	157	26.0	16	17.4	52	34.1	2143	805.6	ND	ND	3.27	2.18	1.05	0.62
1996	1	1	9.2	2.6	3.4	0.7	2.8	0.3	161	34.0	18	13.9	40	29.2	2465	297.2	225	80.3	2.39	1.16	0.82	0.38
	1	18	8.2	2.7	2.4	0.7	2.2	0.3	161	56.5	36	37.6	51	27.8	2663	176.1	190	73.1	1.40	0.56	0.81	0.37
	2	1	8.8	2.6	2.7	0.8	2.2	0.4	160	37.3	8	14.6	41	25.9	2466	275.0	226	52.5	1.77	0.50	0.85	0.36
	2	11	8.4	2.8	3.4	1.6	2.9	1.3	147	41.3	29	24.5	50	25.9	2630	220.7	169	55.7	1.07	0.29	0.77	0.31
1997	1	1	7.3	1.9	2.7	1.0	2.6	0.9	155	33.9	14	14.2	22	23.9	2347	354.4	273	63.8	2.56	1.42	1.51	0.66
	1	18	7.2	1.5	2.6	0.5	2.3	0.4	194	68.6	64	53.3	55	14.5	2995	503.5	197	28.8	1.12	0.50	1.08	0.38
	2	1	6.9	1.7	3.6	1.8	3.1	1.5	156	37.8	13	15.8	17	21.8	2435	351.3	252	62.8	1.68	1.25	1.19	0.83
	2	13	6.5	1.4	2.8	1.9	2.3	0.8	148	38.7	21	12.4	30	20.1	2584	433.5	156	50.6	1.33	1.17	1.06	0.76

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Year	Station	Depth	Total Phosphorus		Total filterable-P		Filterable reactive-P		Total Kjeldahl Nitrogen		Ammonia		Nitrate +Nitrate		Reactive Silicon		Organic Carbon		Chlorophyll <i>a</i>		Phaeophytin <i>a</i>	
		(m)	(ug L ⁻¹)	SD	(ug L ⁻¹)	SD	(ug L ⁻¹)	SD	(ug L ⁻¹)	SD	(ug L ⁻¹)	SD	(ug L ⁻¹)	SD	(ug L ⁻¹)	SD	(ug L ⁻¹)	SD	(ug L ⁻¹)	SD	(ug L ⁻¹)	SD
1998	1	1	9.0	1.7	3.3	0.8	1.9	0.0	193	7.7	21	13.9	38	15.9	2387	73.0	152	118.8	0.10	0.04	0.04	0.02
	1	18	7.5	ND	3.7	ND	1.9	ND	182	ND	25	ND	63	ND	2311	NA	36	ND	0.09	NA	0.03	ND
1999	1	1	17.7	18.3	8.6	10.2	6.8	10.0	247	147.2	36	42.6	124	35.2	2390	431.5	261	122.2	2.94	3.19	0.56	0.35
2000	1	1	9.5	4.3	3.1	1.6	1.8	1.6	57	36.6	19	12.5	72	36.1	ND	ND	ND	ND	2.43	1.46	1.10	0.80
2001	1	1	7.8	5.1	6.4	5.2	8.2	6.7	115	22.2	5	3.6	38	32.5	ND	ND	ND	ND	2.37	0.53	0.30	0.20
2002	1	1	6.4	2.3	4.5	3.1	1.5	0.9	131	15.4	5	2.5	27	18.8	ND	ND	ND	ND	1.36	0.14	0.30	0.20
2003	1	1	6.5	3.0	2.2	0.8	2.1	0.8	ND	ND	6	1.8	54	26.9	ND	ND	ND	ND	1.20	0.20	0.50	0.40
2004	1	1	6.2	3.5	4.3	3.2	2.0	0.7	169	103.8	9	2.8	61	31.5	2764	342.8	ND	ND	1.15	0.18	0.28	0.08
	1	18	5.9	2.3	6.2	8.3	3.5	3.5	ND	ND	19	13.2	80	28.4	2914	277.1	ND	ND	0.70	0.35	0.19	0.11
2005	1	1	11.4	4.4	7.6	3.6	3.6	3.1	161	45.6	4	2.0	41	34.8	2701	243.7	ND	ND	1.60	0.68	0.24	0.11
2006	1	1	7.2	4.3	2.2	1.6	2.3	1.1	97	59.6	7	1.7	28	30.8	ND	ND	ND	ND	1.92	0.32	0.50	0.09
2007	1	1	3.6	0.4	1.1	0.3	1.1	0.6	115	32.4	6	0.7	56	39.5	ND	ND	ND	ND	1.47	0.43	0.21	0.08
Fertilization yrs.																						
1990-2000 Avg		1	7.7	3.4	3.9	2.5	3.0	1.9	155	34.2	13	11.0	55	26.4	2672	275	189	68.6	1.54	1.00	0.59	0.29
1990-2006 Avg		1	7.7	3.3	3.9	2.4	3.0	1.9	151	38.4	11	9.0	48	27.3	2602	314	199	66.4	1.71	0.88	0.59	0.31
1990-2007 Avg		1	7.5	3.2	3.8	2.3	2.9	1.8	149	38.1	10	8.6	49	27.9	2602	314	199	66.4	1.70	0.86	0.57	0.29
Post-fertilization yrs.																						
2001-2006 Avg		1	7.6	3.8	4.5	2.9	3.3	2.2	135	49.3	6	2.4	41	29.2	2732	293	ND	ND	1.60	0.34	0.35	0.18
2001-2007 Avg		1	7.0	3.3	4.0	2.5	3.0	2.0	131	46.5	6	2.2	43	30.7	2732	293	ND	ND	1.58	0.35	0.33	0.17

Appendix A8.–Summary of seasonal mean water chemistry parameters by station and depth for Afognak Lake, 1990-2007.

Year	Station	Depth (m)	Sp. Conductivity		pH		Alkalinity		Turbidity		Color		Calcium		Magnesium		Iron	
			(umhos cm ⁻¹)	SD	(Units)	SD	(mg L ⁻¹)	SD	(NTU)	SD	(Pt units)	SD	(mg L ⁻¹)	SD	(mg L ⁻¹)	SD	(ug L ⁻¹)	SD
1990	1	1	41	1.7	6.8	0.1	6.3	0.5	0.8	0.4	14	3.4	2.9	1.4	0.4	0.3	121	24.3
	1	16	41	1.0	6.7	0.2	6.1	0.6	0.7	0.4	11	2.2	3.2	1.8	0.4	0.3	128	38.7
1991	1	1	38	0.8	6.7	0.1	10.4	7.8	0.9	0.3	13	0.8	2.1	0.3	0.8	0.5	210	31.1
	1	14	38	1.0	6.6	0.2	6.9	0.3	0.9	0.2	16	3.9	1.9	0.1	0.8	0.5	190	45.0
1992	1	1	35	1.2	6.6	0.2	5.8	1.0	0.9	0.5	12	3.4	2.5	0.9	0.6	0.3	157	9.3
	1	24	35	0.5	6.3	0.1	4.9	1.0	0.8	0.6	11	1.5	2.5	1.2	0.6	0.3	162	56.9
1993	1	1	37	1.0	6.6	0.1	7.5	2.7	0.5	0.1	7	7.5	2.2	0.4	1.3	1.1	104	34.9
	1	25	39	4.0	6.4	0.4	7.8	2.1	0.5	0.2	10	10.7	2.6	0.9	0.8	0.1	134	52.0
1994	1	1	39	6.5	6.6	0.2	6.2	2.0	1.1	0.8	5	3.2	2.2	0.9	0.6	0.2	141	44.0
	1	2	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
	1	26	36	0.9	6.3	0.3	6.5	2.5	0.7	0.3	6	4.7	2.2	0.5	0.6	0.2	197	87.7
1995	1	1	60	5.6	6.6	0.2	9.8	1.0	2.0	0.8	11	2.6	3.7	1.4	1.3	0.4	85	45.6
	1	17	60	5.4	6.5	0.2	10.0	1.3	2.3	1.2	9	2.0	3.4	0.5	1.6	0.5	101	33.0
	2	1	58	4.9	6.6	0.2	9.7	1.1	1.9	0.9	11	4.3	3.2	0.3	1.1	0.3	87	55.9
	2	11	58	4.3	6.5	0.2	9.6	1.1	2.0	0.8	10	5.5	3.5	0.4	1.3	0.3	101	53.9
1996	1	1	56	1.5	6.7	0.2	10.5	0.7	1.4	1.0	10	2.5	3.2	0.5	1.3	0.2	54	25.9
	1	18	57	2.7	6.6	0.1	11.2	1.9	1.5	0.7	9	0.5	3.1	0.5	1.1	0.3	72	33.2
	2	1	56	1.4	6.7	0.1	10.7	1.0	1.2	0.6	9	1.3	3.1	0.5	1.1	0.3	54	25.7
	2	11	57	1.1	6.7	0.1	10.7	1.0	1.5	0.6	11	2.6	2.9	0.5	1.5	0.3	89	43.4

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Year	Station	Depth (m)	Sp. Conductivity		pH		Alkalinity		Turbidity		Color		Calcium		Magnesium		Iron	
			(umhos cm ⁻¹)	SD	(Units)	SD	(mg L ⁻¹)	SD	(NTU)	SD	(Pt units)	SD	(mg L ⁻¹)	SD	(mg L ⁻¹)	SD	(ug L ⁻¹)	SD
1997	1	1	53	0.6	7.1	0.2	12.1	1.6	1.1	0.1	9	1.9	3.1	0.4	1.1	0.3	28	16.6
	1	18	58	6.7	6.8	0.2	13.9	3.5	1.7	0.4	10	0.8	2.9	0.5	1.7	1.1	68	37.7
	2	1	53	0.8	7.1	0.1	11.7	0.5	1.0	0.2	11	3.8	3.0	0.3	1.0	0.3	34	17.3
	2	13	53	0.5	7.0	0.1	11.9	0.3	1.3	0.5	10	3.0	2.9	0.3	1.0	0.3	44	25.8
1998	1	1	49	0.6	7.0	0.1	12.6	1.3	1.7	1.2	18	10.7	3.2	0.5	0.8	0.2	26	15.0
	1	18	48	ND	7.0	ND	11.8	ND	2.0	ND	11	ND	3.3	ND	1.0	ND	48	ND
1999	1	1	58	0	6.8	0.2	11.1	0.6	1.6	1.0	11	1.7	3.3	0.3	1.4	0.1	82	43.8
2000	1	1	ND	ND	7.1	0.2	8.7	2.4	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2001	1	1	ND	ND	7.2	0.4	10.1	2.3	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2002	1	1	ND	ND	7.2	0.5	10.1	0.5	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2003	1	1	ND	ND	6.9	0.1	9.8	0.6	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2004	1	1	ND	ND	6.9	0.1	11.4	0.7	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
	1	18	ND	ND	6.8	0.1	10.9	0.7	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2005	1	1	ND	ND	6.8	0.1	10.9	1.1	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2006	1	1	ND	ND	6.8	0.1	11.3	0.9	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2007	1	1	ND	ND	6.8	0.1	10.9	1.2	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
1990-2006 Avg		1	49	2.1	6.8	0.2	9.8	1.5	1.2	0.6	11	3.6	2.9	0.6	1.0	0.3	90.8	30.0
1990-2007 Avg		1	49	2.1	6.8	0.2	9.9	1.5	1.2	0.6	10.7	3.6	2.9	0.6	1.0	0.3	90.8	30.0
2001-2006 Avg		1	ND	ND	7.0	0.2	10.3	1.2	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
2001-2007 Avg		1	ND	ND	6.9	0.2	10.6	1.0	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND

Appendix A9.—Weighted mean zooplankton density, biomass, size by species for station 1 (1987-2007) and station 2 (1988-2007), Afognak Lake.

Station 1	No. Year Samples	<i>Epischura</i>			<i>Diaptomus</i>			<i>Cyclops</i>			<i>Bosmina</i>			<i>Daphnia</i>			<i>Holopedium</i>			TOTALS	
		Density (no/m ²)	Biomass (mg/m ²)	Size (mm)	Density (no/m ²)	Biomass (mg/m ²)	Size (mm)	Density (no/m ²)	Biomass (mg/m ²)	Size (mm)	Density (no/m ²)	Biomass (mg/m ²)	Size (mm)	Density (no/m ²)	Biomass (mg/m ²)	Size (mm)	Density (no/m ²)	Biomass (mg/m ²)	Size (mm)	Density (no/m ²)	Biomass (mg/m ²)
1987	4	28,835	100	0.91	173	1	1.01	4,127	6	0.65	138,370	134	0.33	3,218	4	0.54	2,574	6	0.52	177,297	251
1988	4	22,360	77	0.91	0	0		3,185	5	0.69	106,462	104	0.33	962	2	0.71	1,228	3	0.53	134,197	191
1989	5	16,322	71	0.99	0	0		3,663	5	0.66	69,638	59	0.31	1,778	3	0.64	1,347	3	0.48	92,748	141
1990	7	15,378	60	0.95	7	0	0.90	9,987	16	0.68	155,051	134	0.31	3,392	5	0.61	4,944	9	0.47	188,759	224
1991	6	21,278	102	1.02	265	1	0.79	6,606	12	0.74	208,574	193	0.32	4,089	9	0.72	4,025	8	0.50	244,837	325
1992	7	23,468	104	0.99	485	1	0.88	4,807	8	0.68	106,832	108	0.33	5,513	13	0.74	3,306	6	0.45	144,411	240
1993	7	33,893	127	0.94	76	0	0.83	5,960	11	0.72	240,817	247	0.34	7,689	14	0.66	3,715	8	0.50	292,150	407
1994	8	23,713	66	0.85	1,844	7	0.98	10,231	17	0.69	257,749	256	0.33	9,621	18	0.66	7,271	13	0.48	310,429	377
1995	7	16,758	84	1.04	5,596	16	0.87	24,932	39	0.68	212,768	197	0.32	13,740	22	0.62	1,410	2	0.46	275,204	360
1996	5	42,112	223	1.06	191	0	0.49	11,614	19	0.69	350,806	378	0.34	16,072	44	0.78	2,909	5	0.47	423,704	670
1997	6	14,367	69	1.02	5,520	11	0.75	24,567	41	0.69	81,591	66	0.30	11,720	17	0.58	915	1	0.43	138,679	205
1998	4	15,672	62	0.96	1,088	5	1.05	2,070	3	0.67	169,971	144	0.31	10,881	14	0.56	5,441	8	0.42	205,123	236
1999	4	18,737	78	0.97	5,945	24	0.97	6,688	12	0.71	133,175	130	0.33	9,449	20	0.68	2,495	5	0.46	176,489	269
2000	5	57,643	180	0.88	8,121	44	1.09	10,743	16	0.66	114,297	126	0.35	5,042	9	0.64	1,408	2	0.46	116,722	188
2001	5	30,122	66	0.77	2,548	6	0.79	8,121	10	0.61	40,764	33	0.30	1,253	1	0.49	2,638	4	0.43	85,446	120
2002	4	8,174	21	0.82	1,009	3	0.92	6,380	7	0.56	38,256	36	0.32	2,935	3	0.51	557	1	0.41	57,311	71
2003	4	39,743	73	0.73	3,782	7	0.74	3,185	4	0.62	102,110	85	0.30	1,393	2	0.60	1,194	2	0.48	151,407	173
2004	5	23,206	37	0.69	510	1	0.86	6,374	8	0.62	58,598	52	0.31	11,472	16	0.58	2,771	5	0.48	102,931	119
2005	5	21,369	59	0.84	1,592	4	0.83	8,238	10	0.60	82,409	65	0.30	4,979	7	0.57	2,027	3	0.43	120,614	148
2006	5	29,565	92	0.88	3,450	10	0.85	9,915	20	0.76	76,518	61	0.30	8,408	11	0.56	6,348	11	0.46	134,204	205
2007	5	10,913	24	0.78	2,930	9	0.88	7,718	13	0.70	74,257	66	0.31	3,386	5	0.58	1,730	3	0.47	100,934	120
1987-1989 Avg		22,506	83	0.94	58	0	1.01	3,658	5	0.67	104,823	99	0.32	1,986	3	0.63	1,716	4	0.51	134,747	194
1987-2006 Avg		25,136	88	0.91	2,110	7	0.87	8,570	13	0.67	137,238	130	0.32	6,680	12	0.62	2,926	5	0.47	178,633	246
1987-2007 Avg		24,458	85	0.90	2,149	7	0.87	8,529	13	0.67	134,239	127	0.32	6,523	11	0.62	2,869	5	0.47	174,933	240
2001-2006 Avg		25,363	58	0.79	2,149	5	0.83	7,036	10	0.63	66,443	55	0.31	5,073	7	0.55	2,589	4	0.45	108,652	139
2001-2007 Avg		23,299	53	0.79	2,260	6	0.84	7,133	10	0.64	67,559	57	0.31	4,832	6	0.56	2,466	4	0.45	107,550	136

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Station 2	No. Year Samples	<i>Epischura</i>			<i>Diaptomus</i>			<i>Cyclops</i>			<i>Bosmina</i>			<i>Daphnia</i>			<i>Holopedium</i>			TOTALS	
		Density (no/m ²)	Biomass (mg/m ²)	Size (mm)	Density (no/m ²)	Biomass (mg/m ²)	Size (mm)	Density (no/m ²)	Biomass (mg/m ²)	Size (mm)	Density (no/m ²)	Biomass (mg/m ²)	Size (mm)	Density (no/m ²)	Biomass (mg/m ²)	Size (mm)	Density (no/m ²)	Biomass (mg/m ²)	Size (mm)	Density (no/m ²)	Biomass (mg/m ²)
1988	4	10,656	45	0.98	40	0	1.44	809	1	0.70	108,838	110	0.33	1,405	3	0.65	942	3	0.55	122,690	162
1989	5	10,306	35	0.90	0	0		1,261	2	0.66	48,235	40	0.30	420	1	0.63	553	1	0.46	60,775	79
1990	7	12,610	48	0.94	0	0		3,460	5	0.66	128,277	108	0.31	2,350	4	0.64	4,026	7	0.47	150,723	172
1991	6	19,285	80	0.97	1,274	4	0.89	4,277	8	0.74	154,341	132	0.31	3,347	6	0.65	5,083	10	0.49	187,607	240
1992	7	8,948	34	0.94	144	1	1.00	1,436	2	0.67	82,879	84	0.33	2,521	5	0.70	1,579	3	0.45	97,507	129
1993	7	19,033	70	0.93	773	1	0.69	3,882	5	0.62	175,106	157	0.32	2,570	5	0.67	3,988	7	0.47	205,352	245
1994	8	11,006	40	0.93	783	3	0.91	2,736	4	0.65	125,352	116	0.32	4,321	7	0.64	2,468	4	0.46	146,666	174
1995	7	12,193	44	0.92	1,168	4	0.94	9,054	11	0.61	111,525	98	0.31	8,902	12	0.58	1,152	1	0.4	143,994	170
1996	5	20,892	99	1.02	255	2	1.17	2,930	6	0.77	219,747	239	0.35	4,331	11	0.76	1,571	2	0.46	249,726	359
1997	6	13,677	57	0.97	3,468	7	0.75	3,822	5	0.64	86,060	63	0.29	9,652	13	0.56	924	1	0.41	117,601	146
2004	5	27,192	44	0.70	32	0	0.95	5,125	8	0.66	34,843	27	0.29	2,187	4	0.62	1,624	3	0.44	71,003	84
2005	5	22,282	60	0.83	0	0		2,850	4	0.63	49,992	37	0.29	815	2	0.73	900	1	0.38	76,839	104
2006	5	9,408	14	0.68	510	1	0.78	3,083	5	0.70	44,282	31	0.28	3,571	5	0.59	1,274	2	0.43	62,128	59
2007	5	16,269	63	0.95	1,141	4	0.93	6,693	12	0.71	57,065	49	0.31	934	1	0.55	2,049	4	0.50	84,151	133
1988-2006 Avg		15,673	55	0.92	661	2	0.97	3,470	5	0.67	110,433	101	0.31	3,568	6	0.65	2,068	4	0.45	135,874	172
1988-2007 Avg		15,191	52	0.90	650	2	0.95	3,440	5	0.67	105,344	96	0.31	3,569	6	0.65	2,006	3	0.45	130,201	163